

Uranium Reactive Transport in a Vadose Zone-Aquifer-River System

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Pacific Northwest National Laboratory

Advisory Committee on Nuclear Waste WG Meeting

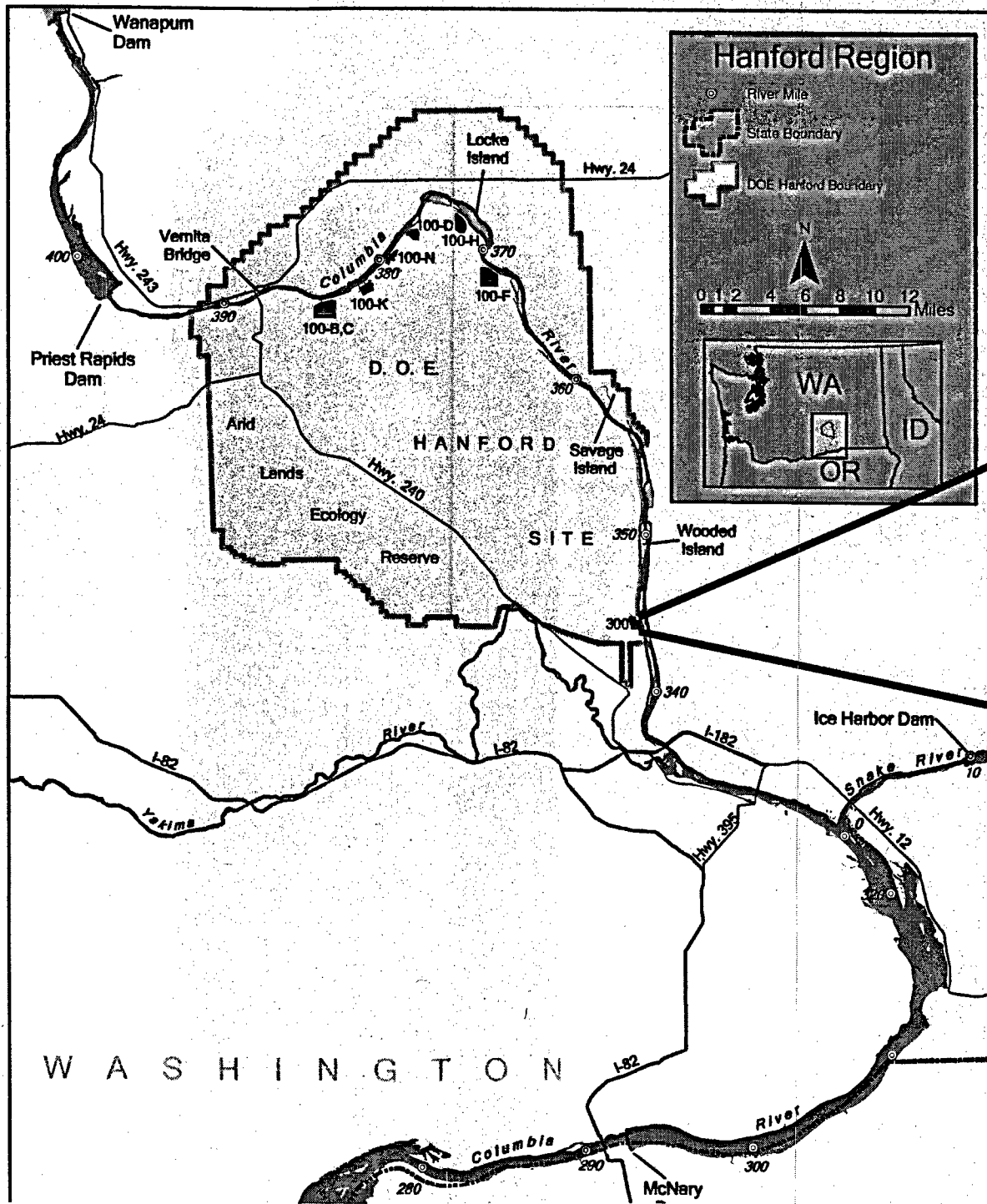
"Integrating Monitoring and Models to Enhance Confidence in Model Results"

September 19-20, 2006

Washington, D.C.

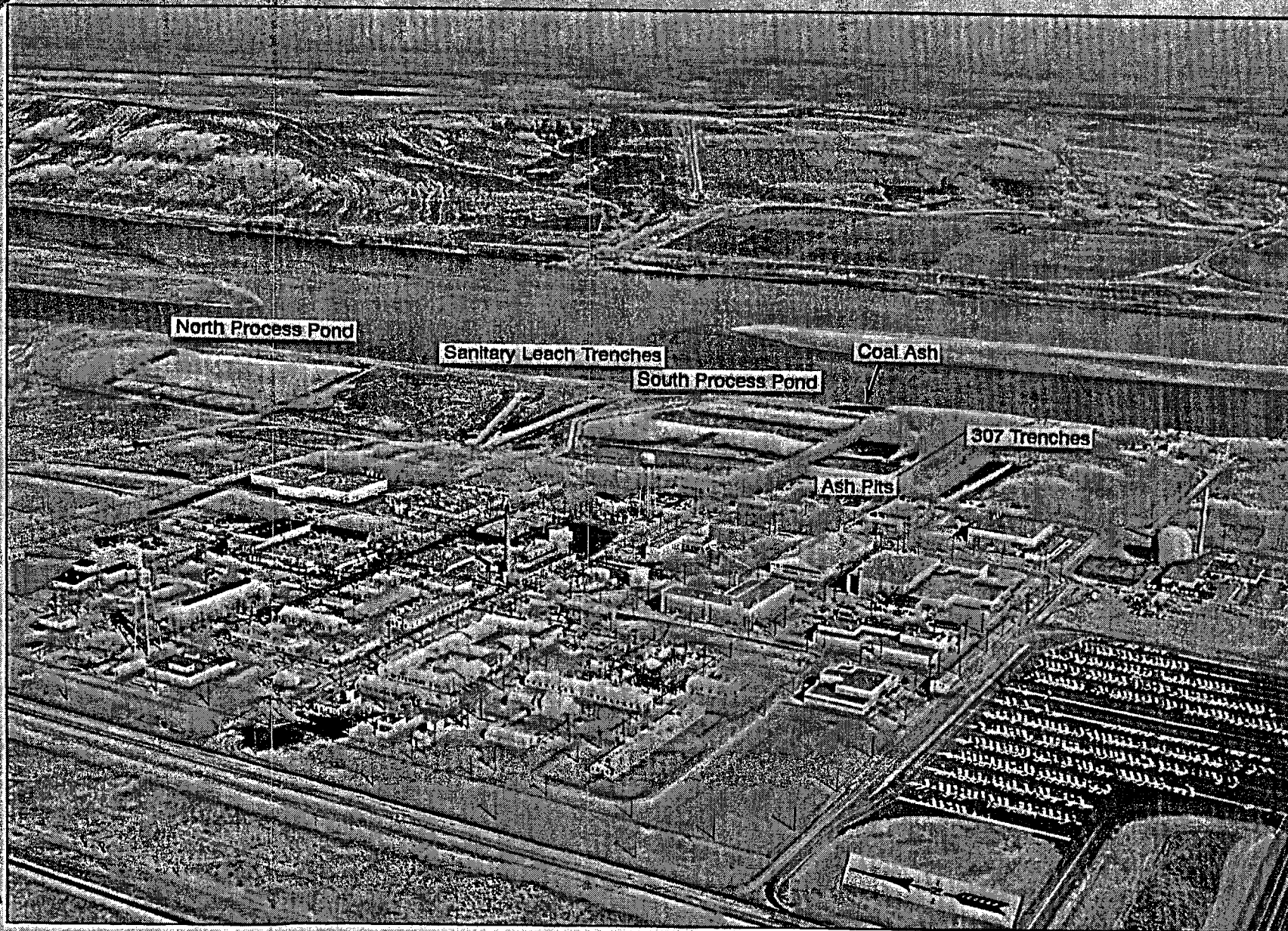
General Themes

- ▶ Monitoring and modeling should be consistent with the scales of the controlling processes
- ▶ Modeling and geophysics can add value
 - sampling scheme design
 - interpretation of monitoring data





Hanford 300 Area in 1962



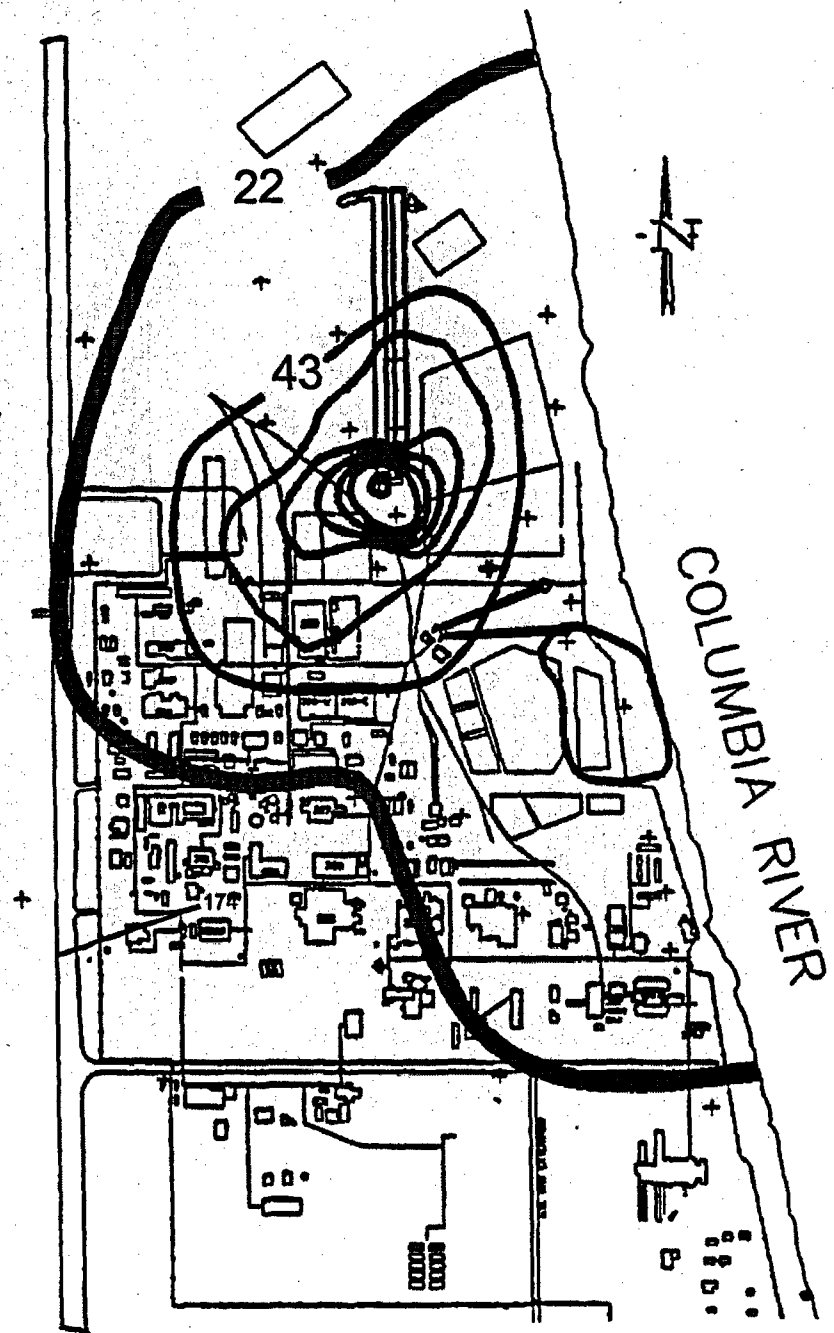
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Operational History and Inventory of 300 Area Process Ponds

- ▶ Fabrication of nuclear fuel elements for the Hanford reactors results in uranium liquid waste streams
 - Unlined waste ponds
 - South Process Pond – 1943 to 1975
 - North Process Pond – 1948 to 1975
 - Unlined process trenches
 - 316-3 Trenches – 1953 to 1963
 - 316-5 Trenches – 1975 to 1994
- ▶ Complex, poorly documented waste disposal history
 - Estimated 70,000 kg of uranium to process ponds
 - 1.55 to 7.57 million liters per day discharge to ponds
 - 10 m to water table
 - Uranium plumes tracked in the Columbia River in 1957 and 1962

1990 Uranium Plume

- ▶ Large area exceeding drinking water standard
- ▶ Hot spot at south end of 316-5 trenches at high river stage
- ▶ 1991 Expedited Response Action
 - Remove contaminated soils from process trenches
 - End discharge of uranium to process trenches
- ▶ 1993 groundwater flow and uranium transport analysis predicted cleanup to < 20 ug/L in 3 to 10 years



1990 Uranium (ug/L)

1993 Conceptual Model

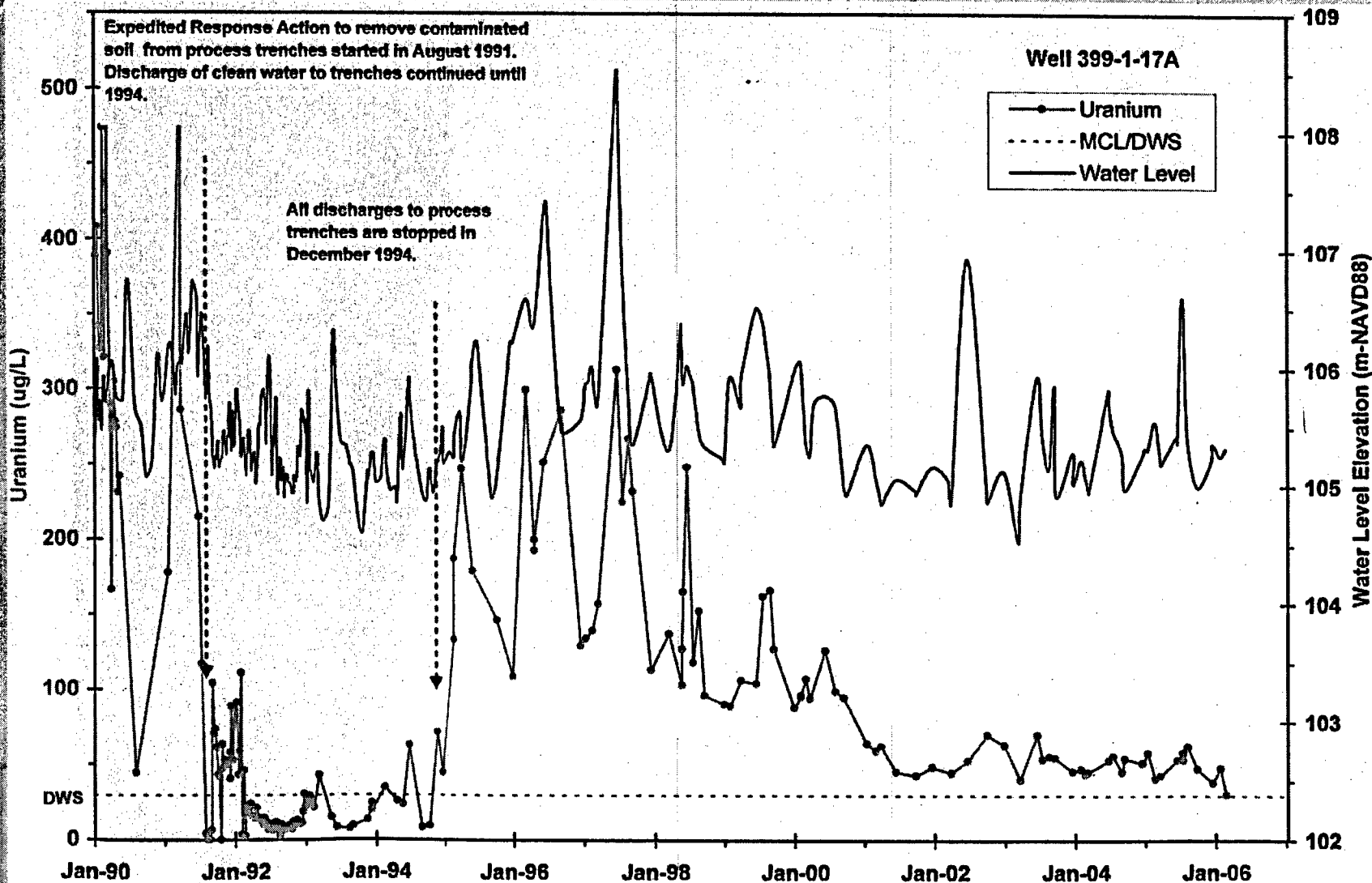
Modeling Assumptions in Phase I Remedial Investigation (1994)

- ▶ 3-D saturated unconfined aquifer; **vadose zone not modeled**
 - Spatially distributed hydraulic conductivity (4 hydrofacies types)
 - Flow field driven by **monthly** changes in river stage fluctuations
 - Uranium mobility controlled by **constant K_d**
 - Natural flushing predicted to largely decrease U to < 20 ug/L by 2018

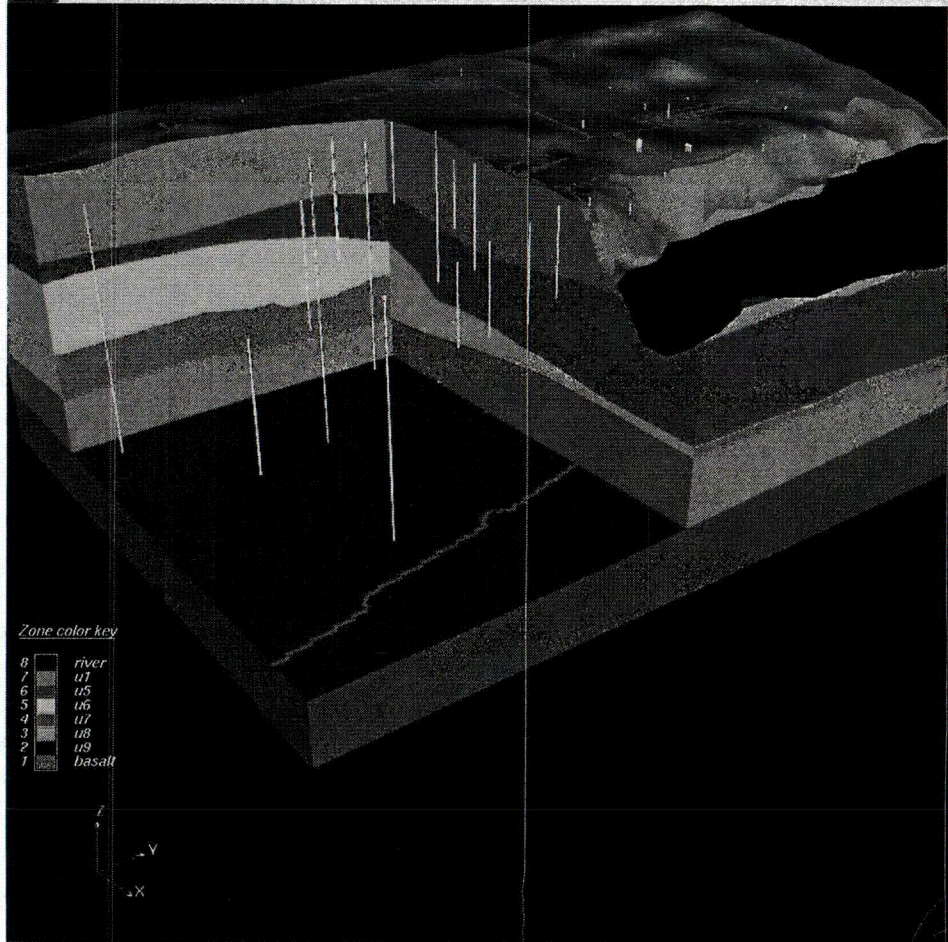
Prediction Update for $U < 20$ ug/L in RI/FS (1995)

- ▶ "Refinement" of Phase I RI estimate: **3 to 10 years** from late 1993 to meet standard
- ▶ Analytical model assumptions
 - Steady-state saturated flow
 - Constant hydraulic conductivity: 1830 m/day
 - Constant hydraulic gradient: 5×10^{-4}
 - 500 m travel distance from process trenches to Columbia River
 - Uranium mobility controlled by "best estimate" constant $K_d \sim 1$ to 2 mL/g
- ▶ No interaction between aquifer and river
- ▶ No interaction between aquifer and vadose zone

Aquifer Water Levels and Uranium Concentrations



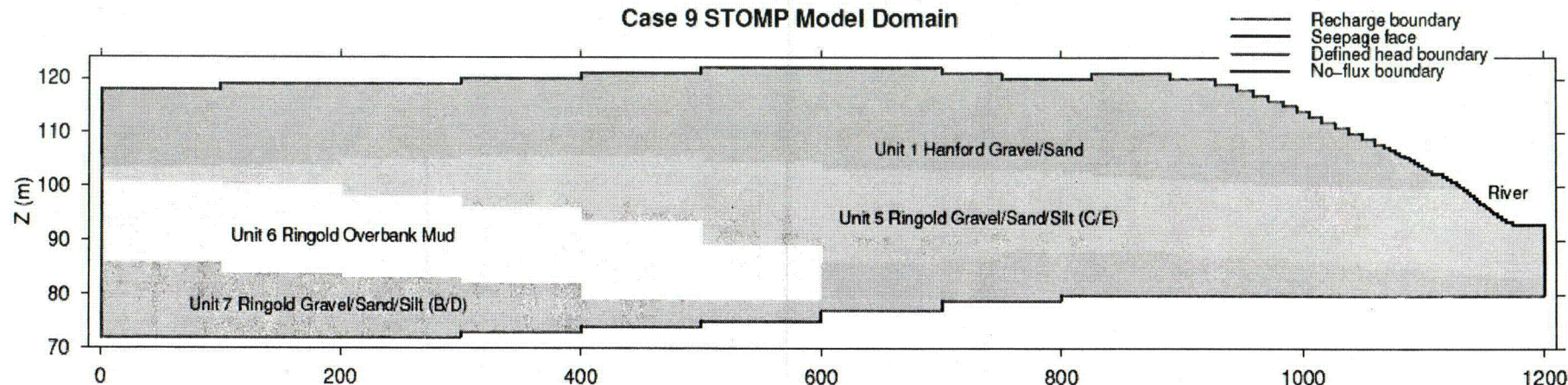
Flow and Transport: Vadose Zone – Aquifer – River System



- ▶ 2-D and 3-D modeling
- ▶ Most current hydrogeology
- ▶ Flow and transport driven by hourly river stage fluctuations
- ▶ Investigate dynamics of riverbank storage and fluxes across aquifer - river interface
- ▶ Investigate release of uranium from contaminated vadose zone sediments due to water table fluctuations

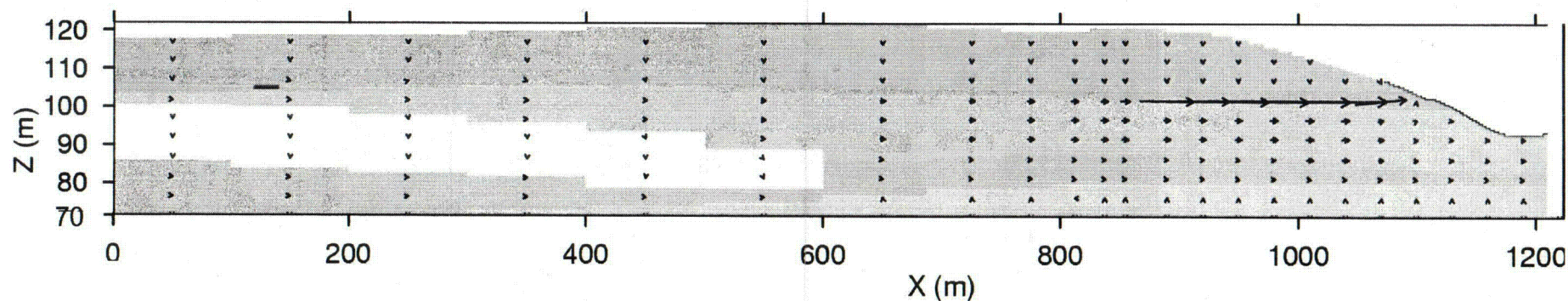
Flow and Transport: Vadose Zone – Aquifer – River System

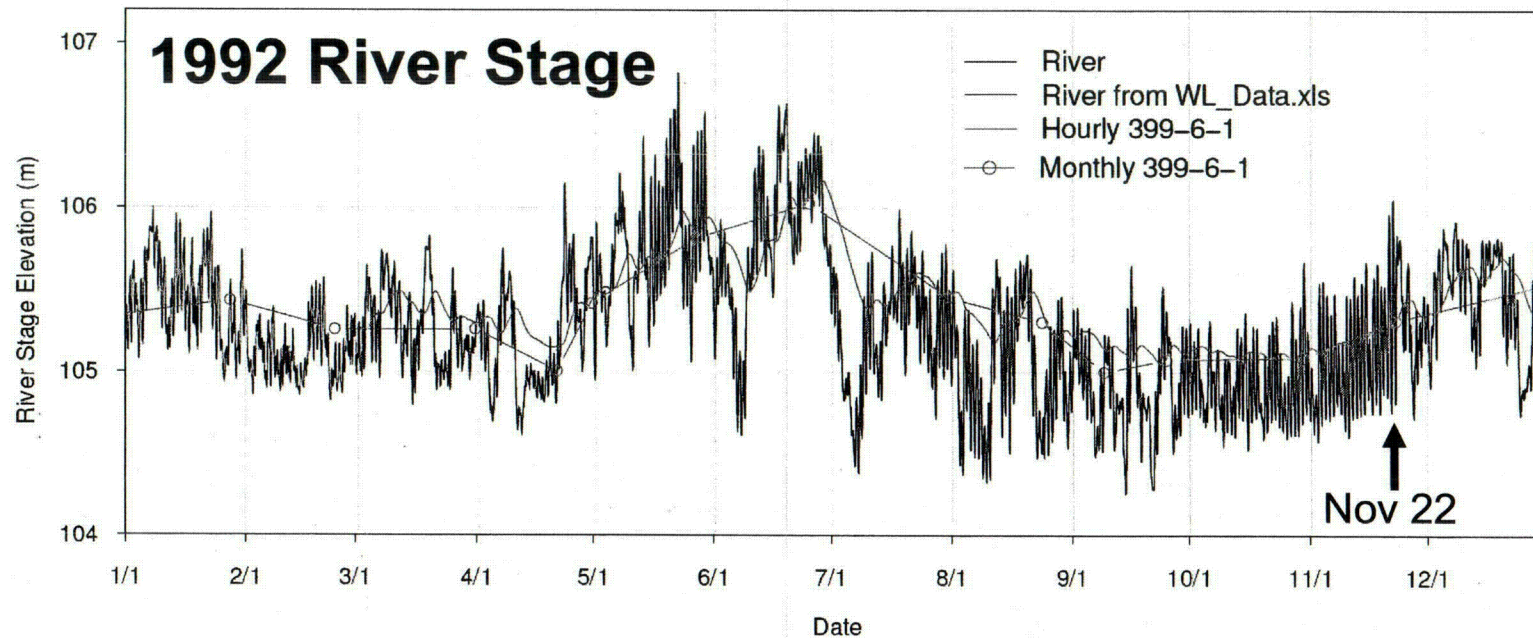
Case 9 STOMP Model Domain



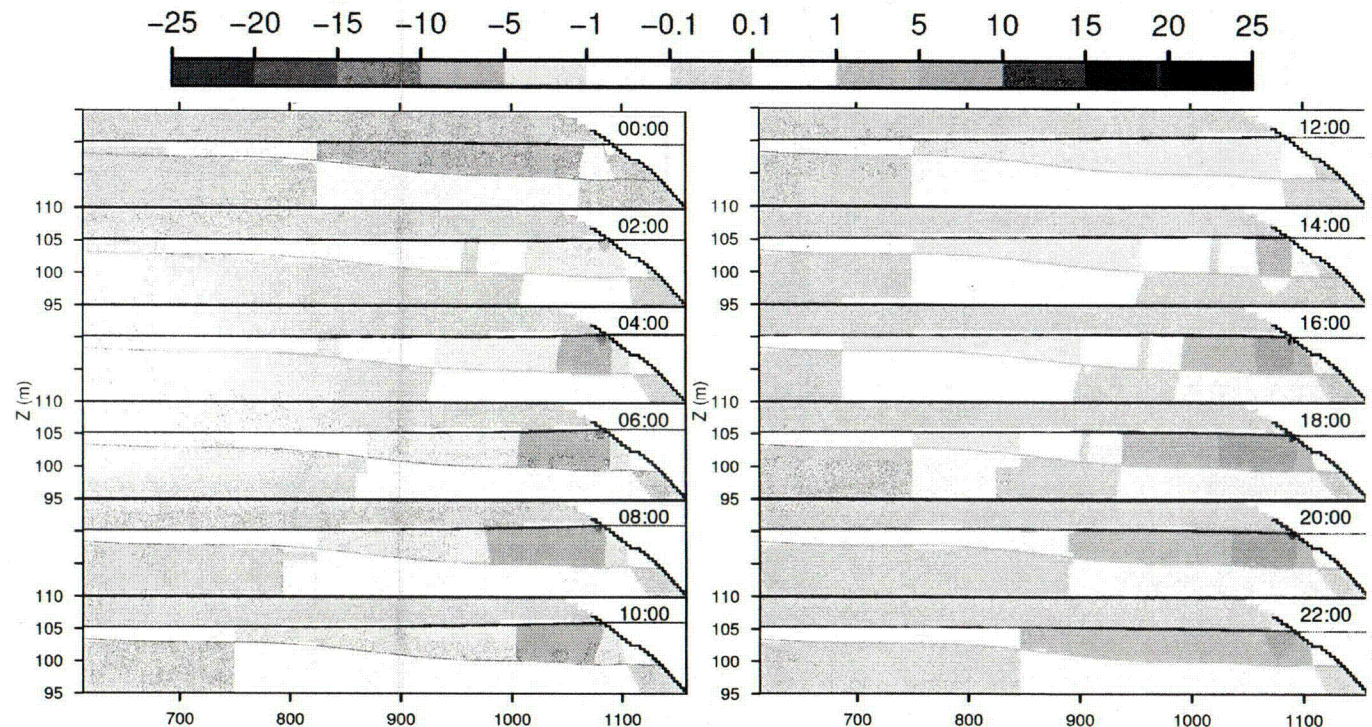
Material	$K_l^{(a)}$ (m d^{-1})	$\phi^{(b)}$	$\psi^{(c)}$ (cm)	$\lambda^{(d)}$	$\theta_r^{(e)}$
U1, Hanford Gravel/Sand	1500	0.25	23.04	0.7465	0.1471
U5, Ringold Gravel/Sand/Silt (C/E)	150	0.18	71.31	0.5193	0.1299
U6, Ringold Overbank Mud	0.01	0.18	71.31	0.5193	0.1299
U7, Ringold Gravel/Sand/Silt (B/D)	43	0.18	71.31	0.5193	0.1299
U8, Lower Ringold Mud	5e-5	0.18	71.31	0.5193	0.1299
Basalt	5e-5	0.18	71.31	0.5193	0.1299

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X-Direction Groundwater Flux (m/d) (Nov 22, 1992)

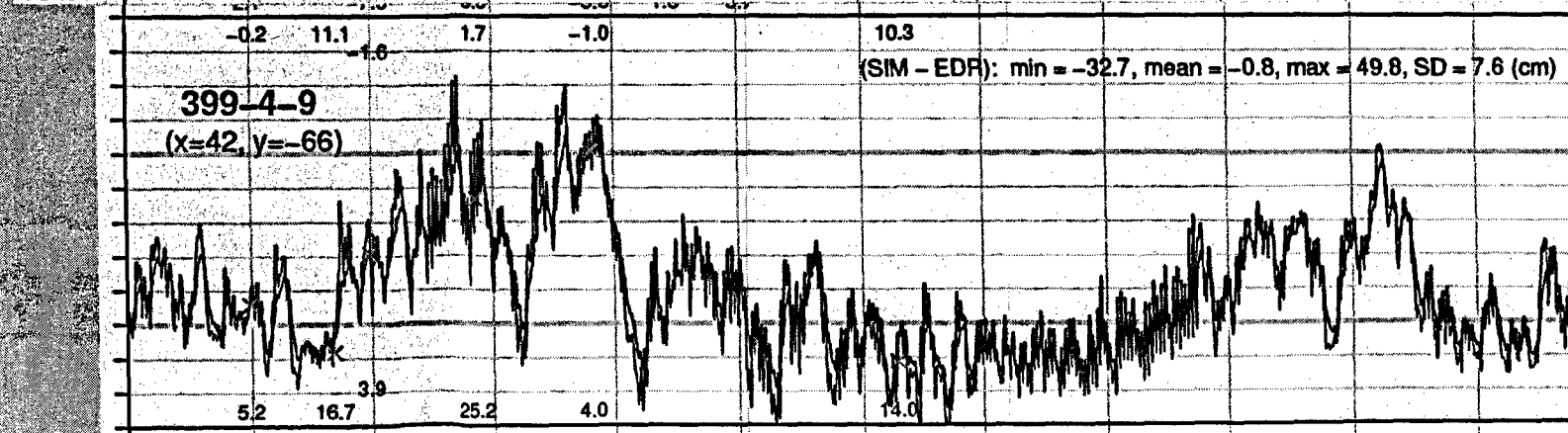
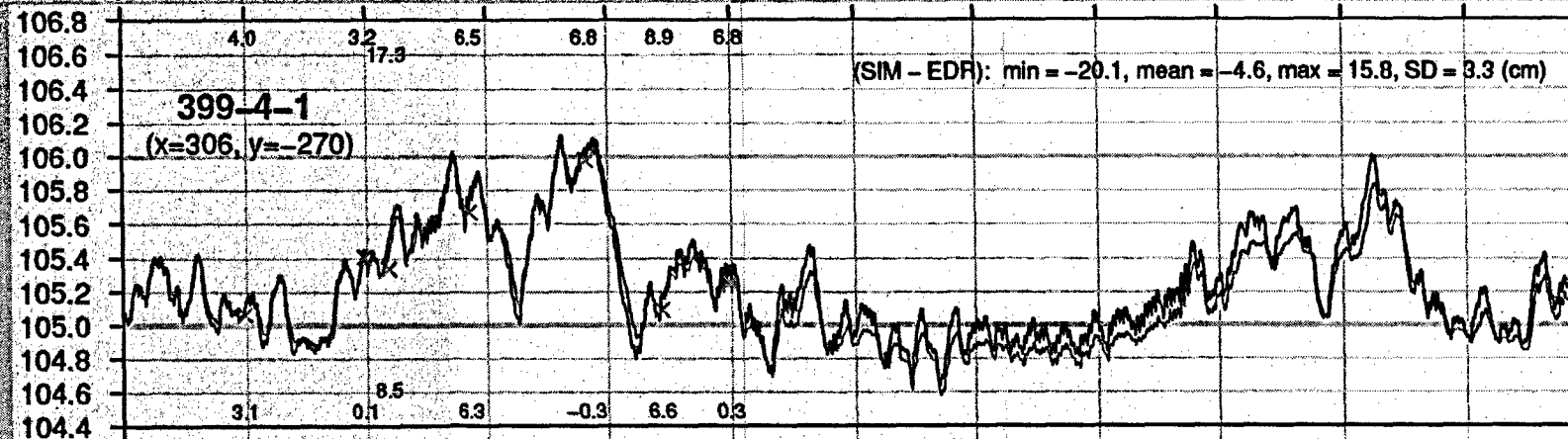
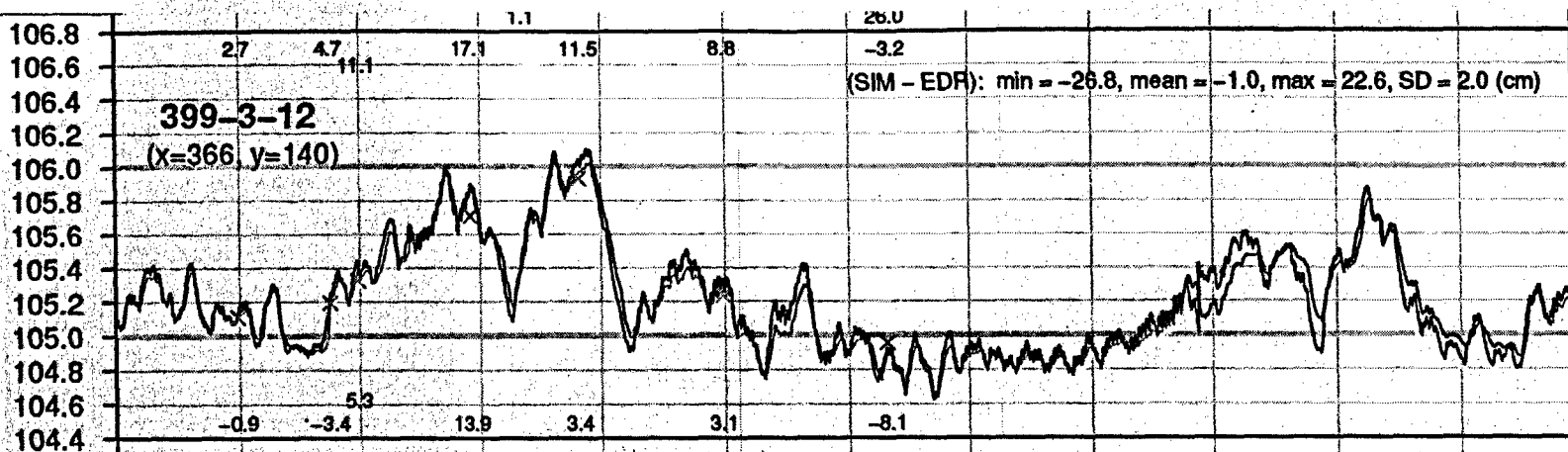


— STOMP, sp. interp. — Edrington × HEIS

Green numbers are (SIM - HEIS) in cm

Blue numbers are (EDR - HEIS) in cm

Hydraulic Head (m)

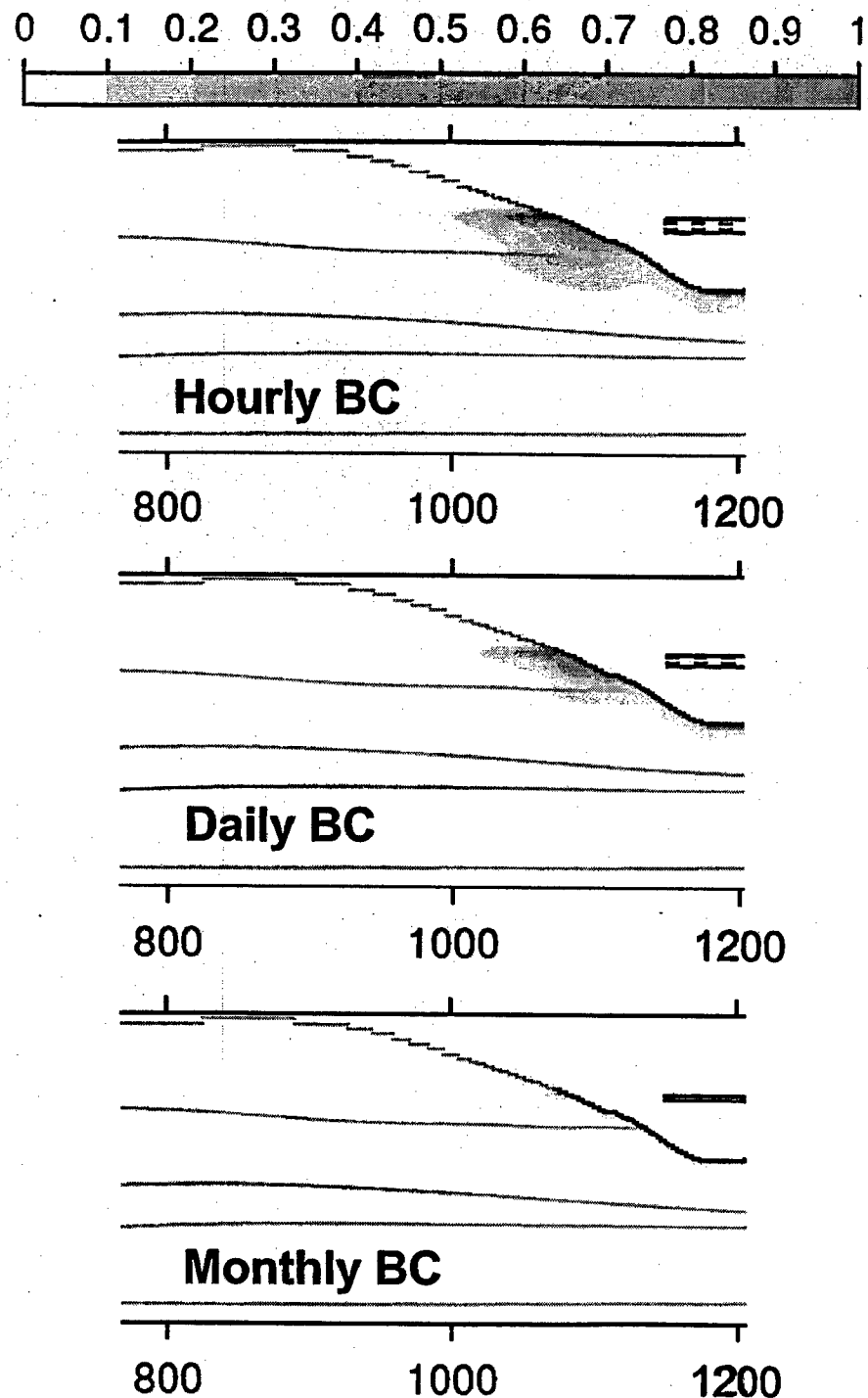


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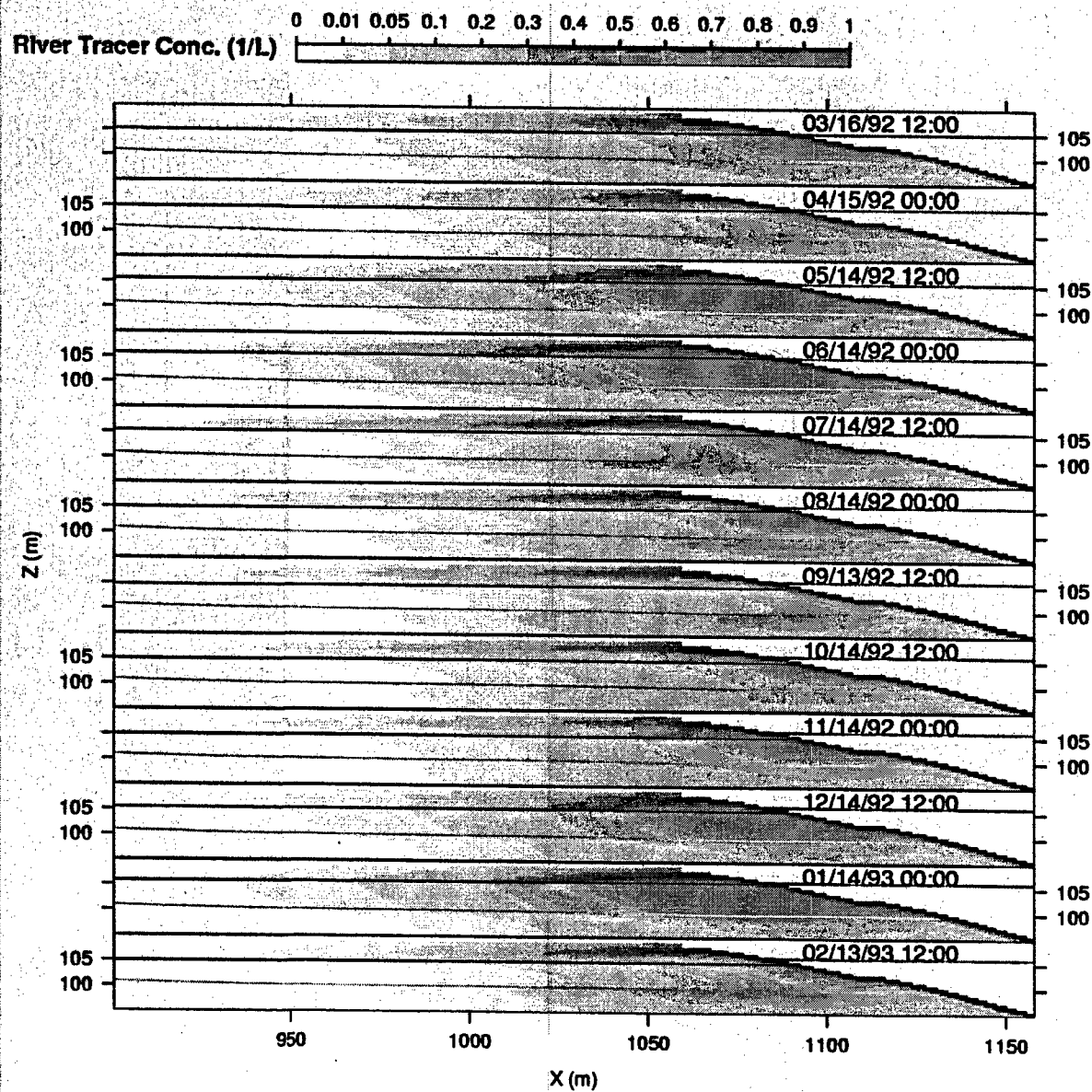
Aquifer-River Mixing

- ▶ “Normal” mixing zone extends ~150 m inland
- ▶ Averaging river stage fluctuations over daily period reduces size of mixing zone
- ▶ Monthly average essentially eliminates mixing with river water

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Seasonal Variation in Mixing Zone

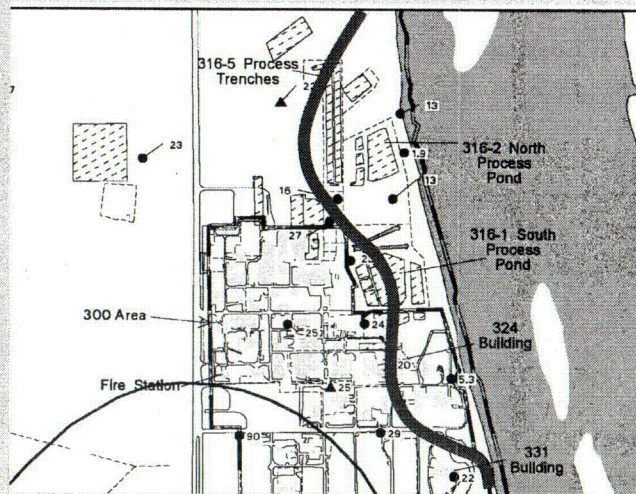


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Aquifer-River Solution Chemistry

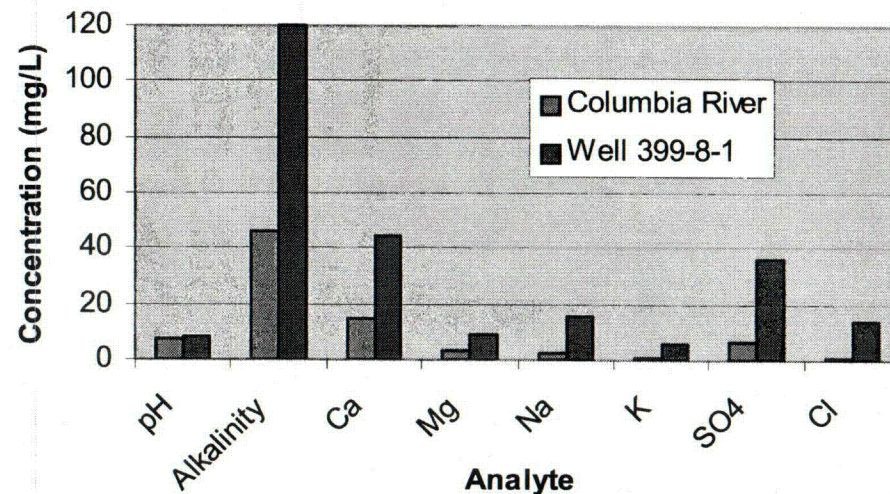
- ▶ River water influx occurs during high stage
- ▶ Prolonged seasonal high stage period allows mixing in aquifer with river water
- ▶ Significant differences in solution chemistry
- ▶ High pore velocity observed: 10 m/d pore velocity (Cline et al. 1985)



2002 Nitrate Concentration

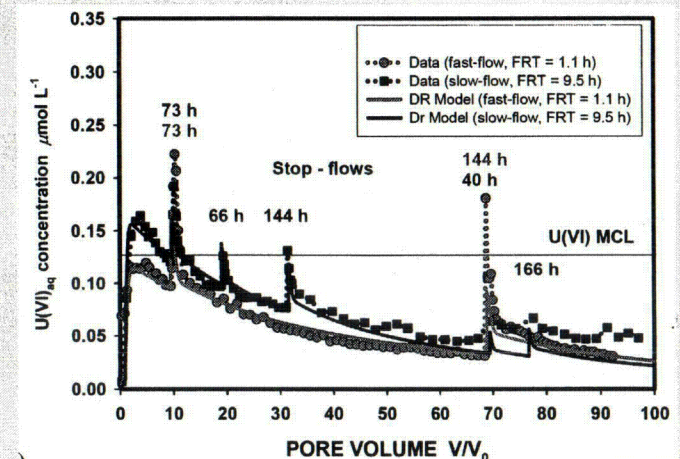
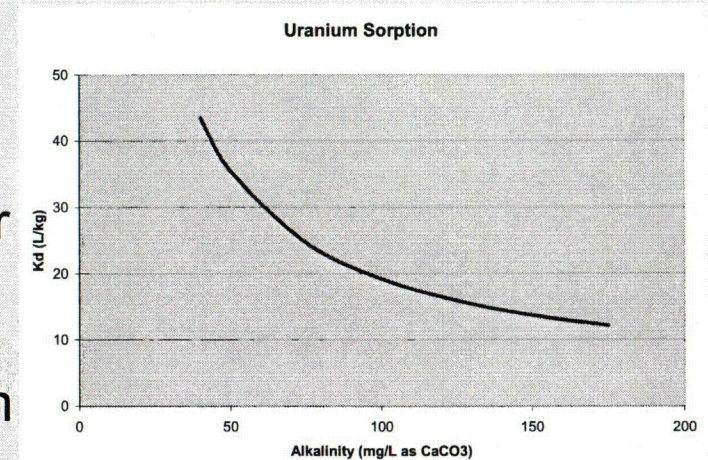
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River versus Aquifer Water Chemistry



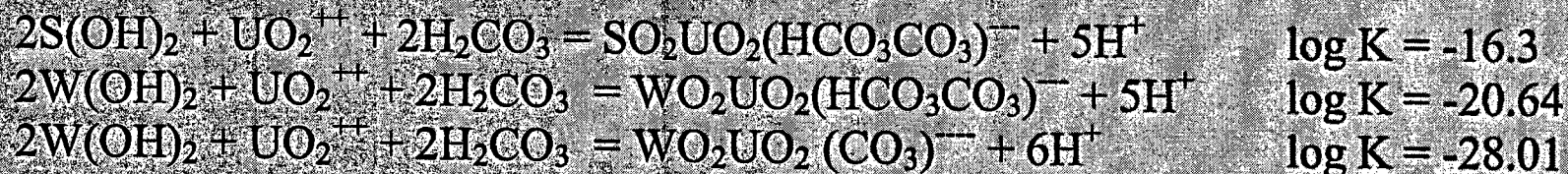
Uranium Geochemistry

- ▶ Constant K_d not consistent with experimental observations
 - Uranium sorption varies strongly with transition between aquifer and river water chemistries (e.g., U, Ca, pH, alkalinity concentrations)
 - Rate-limited uranium sorption identified in column experiments with flow rates consistent with field observations
- ▶ Key Issues
 - Uranium leaching from contaminated vadose zone sediments by water table fluctuations
 - Changing uranium geochemistry during mixing and exchange of river and groundwater

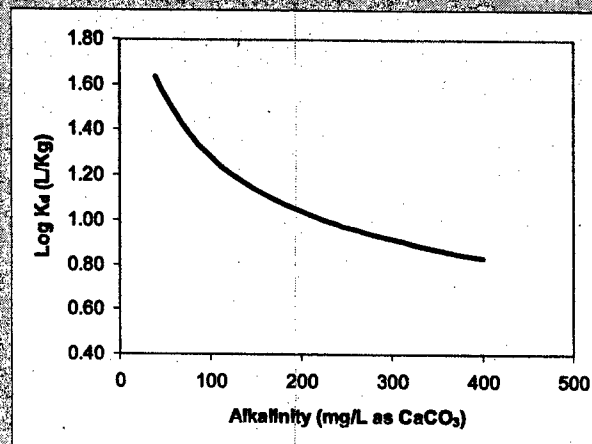


Uranium Batch Studies

- ▶ Performed by Jim Davis, USGS
- ▶ Preliminary three-reaction generalized composite surface complexation model
 - accounts for bicarbonate concentration, sediment surface area, and aqueous U(VI) complexation (21 reactions)
 - 1 strong site and 2 weak site reactions:



- The total site density of 3.84 $\mu\text{moles/m}^2$ is comprised of a strong site density of 1.344 nmoles/m^2 and a weak site density of 3.839 $\mu\text{moles/m}^2$
- ▶ Kd vs alkalinity



Uranium Aqueous Complexes

Reaction	$\log \beta^* (I = 0)^a)$
$\text{UO}_2^{2+} + \text{H}_2\text{O} \rightleftharpoons \text{UO}_2\text{OH}^+ + \text{H}^+$	-5.25
$\text{UO}_2^{2+} + 2\text{H}_2\text{O} \rightleftharpoons \text{UO}_2(\text{OH})_2^{\text{aq}} + 2\text{H}^+$	-12.15
$\text{UO}_2^{2+} + 3\text{H}_2\text{O} \rightleftharpoons \text{UO}_2(\text{OH})_3^- + 3\text{H}^+$	-20.25
$\text{UO}_2^{2+} + 4\text{H}_2\text{O} \rightleftharpoons \text{UO}_2(\text{OH})_4^{2-} + 4\text{H}^+$	-32.4
$2\text{UO}_2^{2+} + \text{H}_2\text{O} \rightleftharpoons (\text{UO}_2)_2\text{OH}^{3+} + \text{H}^+$	-2.70
$2\text{UO}_2^{2+} + 2\text{H}_2\text{O} \rightleftharpoons (\text{UO}_2)_2(\text{OH})_2^{2+} + 2\text{H}^+$	-5.62
$3\text{UO}_2^{2+} + 4\text{H}_2\text{O} \rightleftharpoons (\text{UO}_2)_3(\text{OH})_4^{2+} + 4\text{H}^+$	-11.90
$3\text{UO}_2^{2+} + 5\text{H}_2\text{O} \rightleftharpoons (\text{UO}_2)_3(\text{OH})_5^+ + 5\text{H}^+$	-15.55
$3\text{UO}_2^{2+} + 7\text{H}_2\text{O} \rightleftharpoons (\text{UO}_2)_3(\text{OH})_7^- + 7\text{H}^+$	-32.20
$4\text{UO}_2^{2+} + 7\text{H}_2\text{O} \rightleftharpoons (\text{UO}_2)_4(\text{OH})_7^+ + 7\text{H}^+$	-21.9
$\text{UO}_2^{2+} + \text{CO}_3^{2-} \rightleftharpoons \text{UO}_2\text{CO}_3(\text{aq})$	9.94
$\text{UO}_2^{2+} + 2\text{CO}_3^{2-} \rightleftharpoons \text{UO}_2(\text{CO}_3)_2^{2-}$	16.61
$\text{UO}_2^{2+} + 3\text{CO}_3^{2-} \rightleftharpoons \text{UO}_2(\text{CO}_3)_3^{4-}$	21.84
$2\text{UO}_2^{2+} + \text{CO}_3^{2-} + 3\text{H}_2\text{O} \rightleftharpoons (\text{UO}_2)_2\text{CO}_3(\text{OH})_3^- + 3\text{H}^+$	-0.855
$\text{Ca}^{2+} + \text{UO}_2^{2+} + 3\text{CO}_3^{2-} \rightleftharpoons \text{CaUO}_2(\text{CO}_3)_3^{2-}$	25.4 ^{b)}
$2\text{Ca}^{2+} + \text{UO}_2^{2+} + 3\text{CO}_3^{2-} \rightleftharpoons \text{Ca}_2\text{UO}_2(\text{CO}_3)_3(\text{aq})$	30.55 ^{b)}
$\text{UO}_2^{2+} + \text{NO}_3^- \rightleftharpoons \text{UO}_2\text{NO}_3^+$	0.3
$\text{UO}_2^{2+} + \text{Cl}^- \rightleftharpoons \text{UO}_2\text{Cl}^+$	0.17
$\text{UO}_2^{2+} + 2\text{Cl}^- \rightleftharpoons \text{UO}_2\text{Cl}_2(\text{aq})$	-1.1
$\text{UO}_2^{2+} + \text{SO}_4^{2-} \rightleftharpoons \text{UO}_2\text{SO}_4(\text{aq})$	3.15
$\text{UO}_2^{2+} + 2\text{SO}_4^{2-} \rightleftharpoons \text{UO}_2(\text{SO}_4)_2^{2-}$	4.14

a) Values from Guillaumont *et al.* (2003), unless otherwise indicated.

b) Bernhard *et al.* (2001)

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Uranium Column Studies

- ▶ Performed by Chongxuan Liu, PNNL
- ▶ Saturated column experiments exhibit uranium kinetics
- ▶ Multisite model with different reaction rates and/or diffusive mass transfer rates (Culver et al., 1997):

$$\frac{\partial S}{\partial t} = \sum_{i=1}^N \frac{\partial S_i}{\partial t}; \quad \frac{\partial S_i}{\partial t} = \alpha_i [f_i(\alpha_i) K_d^i C - S_i]$$

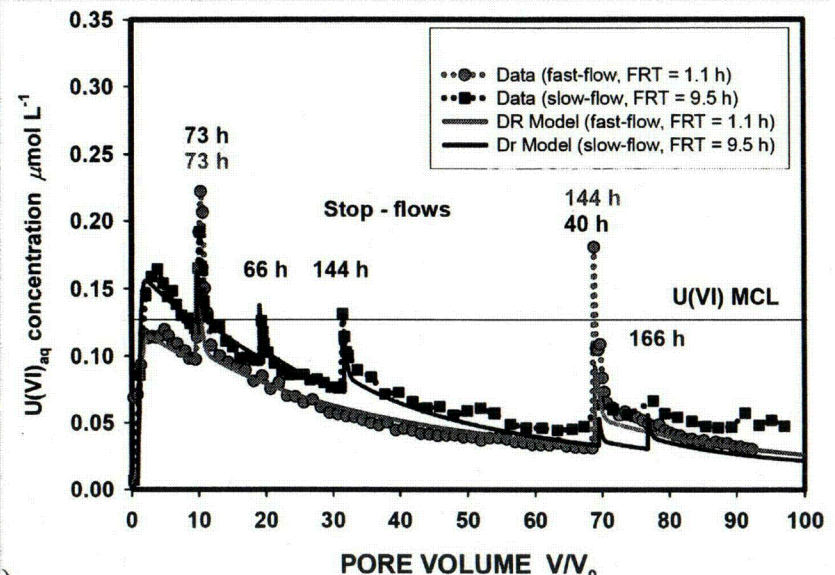
- ▶ Distributed rate parameters were assumed to follow the Gamma statistical distribution (two parameters):

$$f_i(\alpha_i) = \int_{\alpha_i}^{\alpha_i + \Delta \alpha_i} \frac{\beta^{-\eta} \tau^{\eta-1}}{\Gamma(\eta)} \exp\left(-\frac{\tau}{\beta}\right) d\tau$$

- ▶ The multisite kinetic model was integrated into the advection-dispersion equation:

$$\theta \frac{\partial C}{\partial t} + \rho_b \frac{\partial S}{\partial t} = \theta D \frac{\partial^2 C}{\partial x^2} - \theta v \frac{\partial C}{\partial x}$$

U(VI) Breakthrough Curve



Field-Based Reactive Transport Modeling

Account for full sediment size distribution

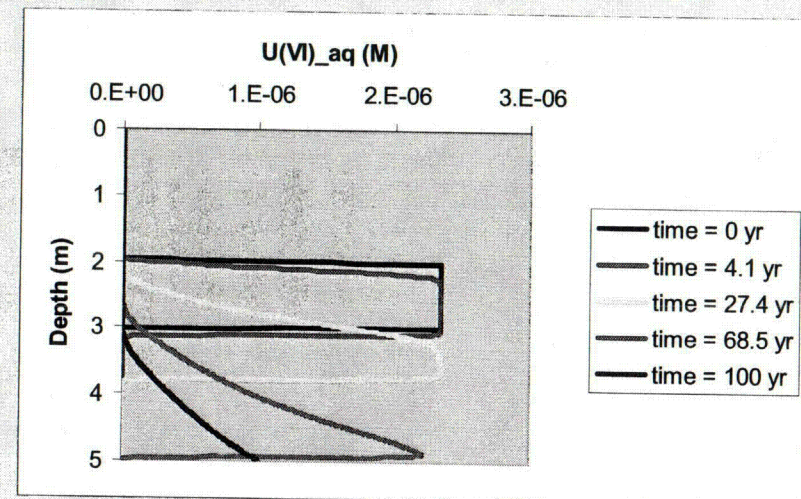
- < 2 mm size fraction in the lab studies
 - Specific surface area: 27.2 m²/g
 - 8% of total sediment
- Preliminary assumption: gravels are unreactive
 - apportion 8% of the 2.06 kg/L field bulk density for surface complexation

Size (mm)	Mass Distribution (%)
Cobbles	
>12.5	74.5
2.0 – 12.5	17.2
Sand	
1.0 – 2.0	2.64
0.5-1.0	2.34
0.25 – 0.5	0.78
0.149 – 0.25	0.33
0.106 – 0.149	0.19
0.053 – 0.106	0.20
Silt + Clay	
<0.053	1.78

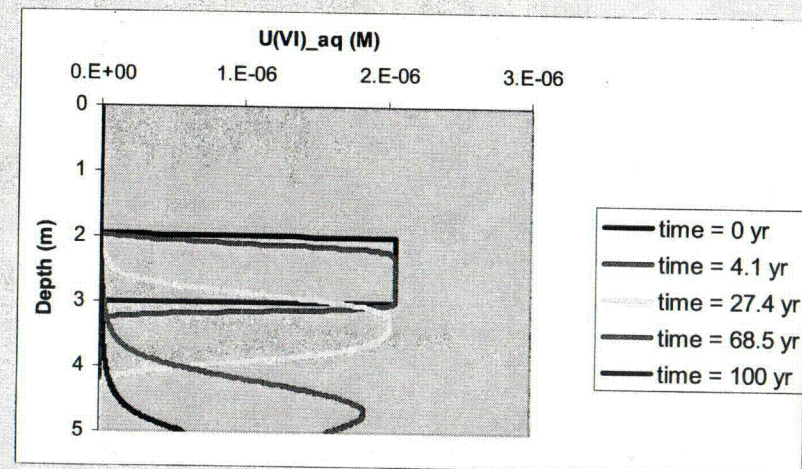
Unsaturated Flow Model Parameters	Value	Units
Horizontal Hydraulic Conductivity	1500	m/d
Vertical Hydraulic Conductivity	150	m/d
Air entry pressure	23.04	cm
Brooks-Corey λ	0.7465	
Residual Saturation	0.1471	
Relative Permeability Method	Burdine	
Porosity	0.25	
Bulk Density	2.06	Kg/L
Recharge Rate	60	mm/yr
Calculated Water Content	0.08	

1-D Unsaturated Reactive Transport Simulation

- ▶ 1-D reactive transport simulation
 - 60 mm/yr recharge results in 0.75 m/yr pore velocity
 - 5 m of vadose zone
 - 1 m of contaminated sediment in the middle
 - 30 nM/g U contaminated zone
- ▶ GC-SCM
 - Sorption front requires over 30 years to move 1 m
 - $K_d = 12.4$ L/kg for this solution chemistry
 - Lowest sediment contamination level results in U(VI) above MCL (0.126 μ M)
- ▶ Multisite kinetic model
 - Very similar to GC-SCM result
 - $K_d = 14$ similar to the GC-SCM
 - impact of kinetics largely minimized by long transport time scales



Generalized Composite SCM



Multisite Kinetic Model

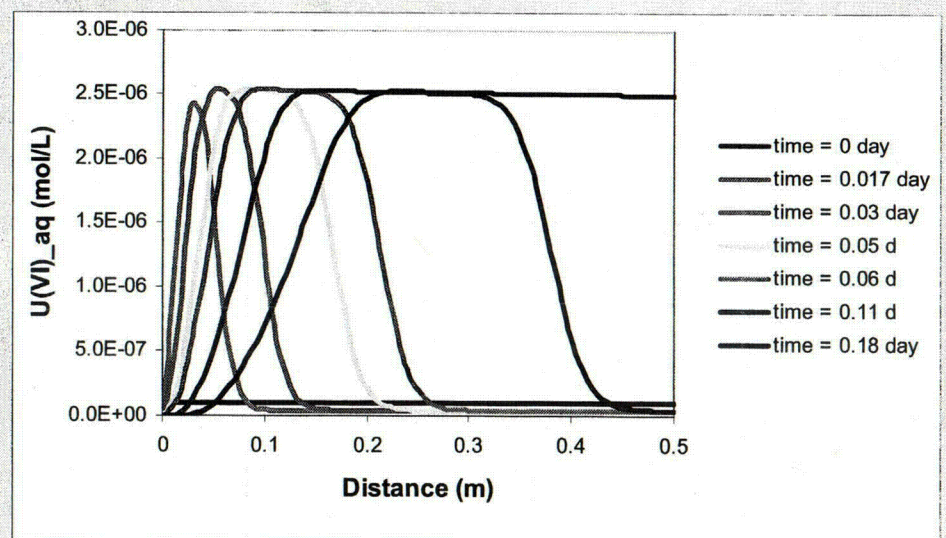
1-D Aquifer-River Interactions

Adapt GC-SCM for the situation where the solution chemistry changes from river water to groundwater

- 1.4 m/d groundwater
- 30 nM/g U-contaminated sediments
- Initial equilibrium with river water
 - $5.76\text{E-}8$ M aqueous U
 - Intrinsic $K_d > 500$ L/kg
- After influx of groundwater
 - Aqueous U is $2.50\text{E-}6$ M
 - Intrinsic $K_d = 13.5$ L/kg

Solution Chemistry

Components	River water (USGS 6/1/2000)	1988 Well 399-8-3
pH	7.1	7.7
HCO ₃ ⁻	9.18e-4 M	2.66e-3 M
K ⁺	1.75e-5	1.50e-4
NO ₃ ⁻	8.55e-6	1.73e-4
Sr ⁺⁺	1.23e-6	0
Na ⁺	1.00e-4	9.87e-4
Ca ⁺⁺	3.74e-4	1.10e-3
Mg ⁺⁺	1.48e-4	4.10e-4
Cl ⁻	3.10e-5	2.75e-3
SO ₄ ⁻⁻	7.08e-5	3.25e-4



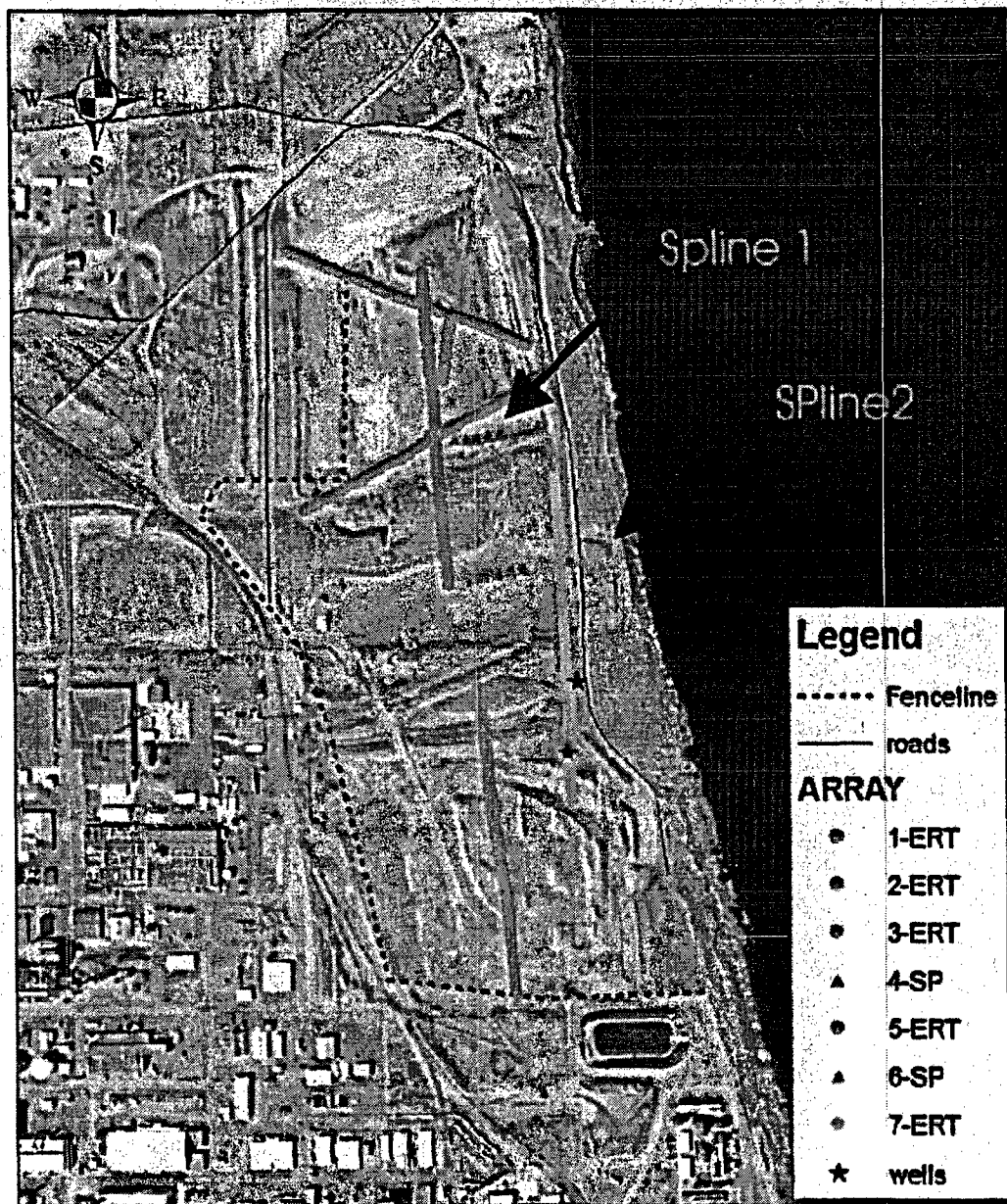
Modeling Summary

Interaction between hourly river stage dynamics, highly transmissive and heterogeneous sediments, and spatially variable uranium create field situation more complex than 1993 conceptual model

- Lower vadose zone uranium accessed by high river stage
- Diurnal cycling of high pore velocities
- Mixing zone of aquifer and river water chemistries
 - dictated by river forcing and hydraulic conductivity
 - sensitive to temporal resolution
 - implications for uranium mobility

► Work in progress

- Ongoing limited field investigation (LFI): sediment cores for detailed analysis, geophysical logging to map uranium distribution
- Laboratory studies provide framework for understanding uranium mobility
 - Solution chemistry
 - Kinetics
- Field-scale studies identify large-scale transport context for understanding uranium fate



Geophysical Characterization

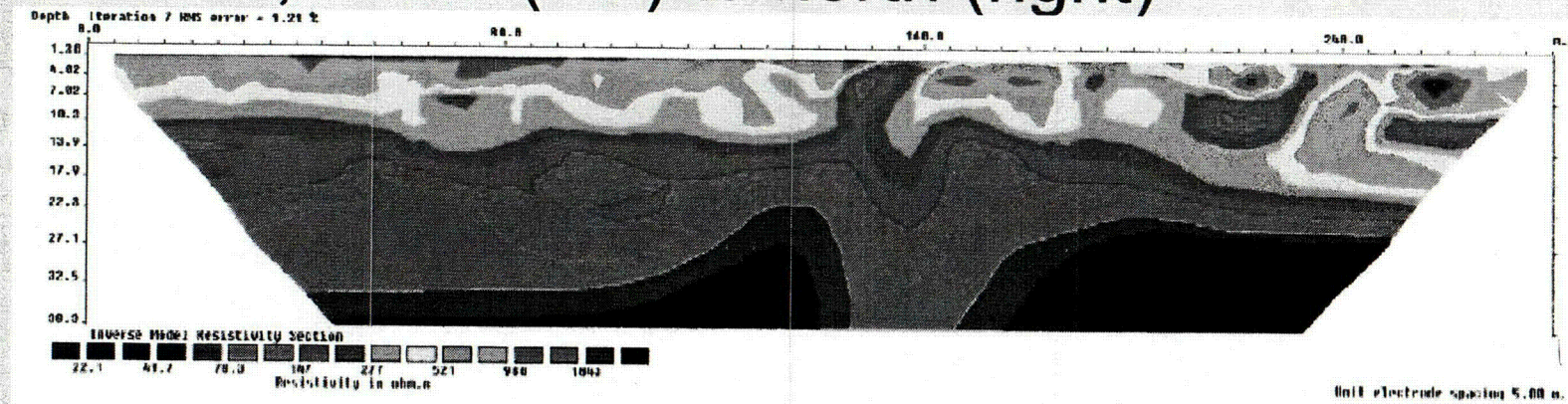
- ▶ Performed by Andy Ward (PNNL) and Roelof Versteeg (INL)
- ▶ March 2006: successful test of single ERT and SP lines
- ▶ August 2006: Full grids of SP, ERT-Complex Resistivity electrodes

Preliminary Resistivity Plots

North Process Pond

(high resistance ~ coarse unsaturated sediment)

► Line 2, south (left) to north (right)



► Line 3, west (left) to east (right)

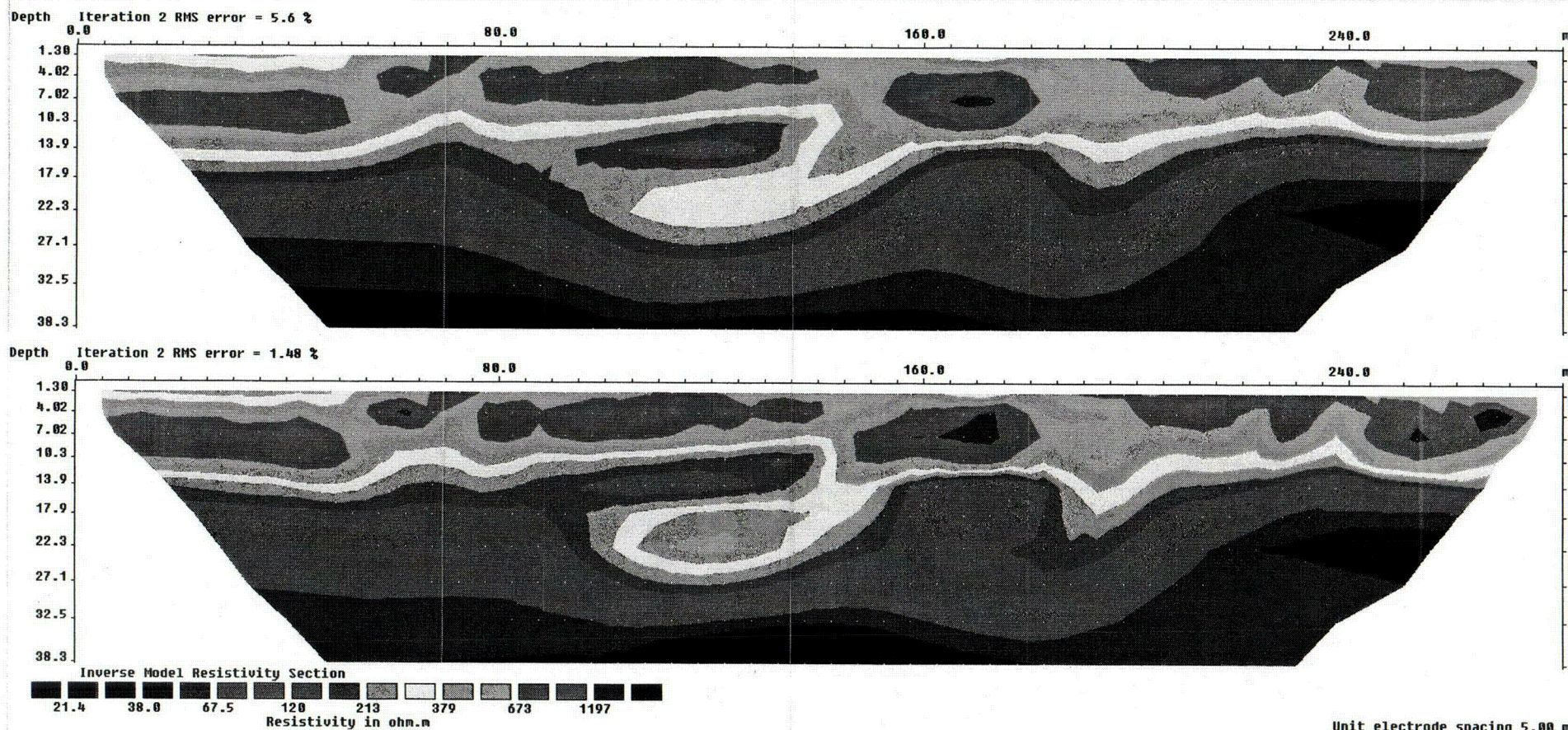


Time-Lapse ERT

Parallel to Shoreline between Process Ponds

► Line 7, south (left) to north (right)

► 3-Day time series



First Grid Deployment

- ▶ 120 SP and 60 ERT electrodes
- ▶ 30 m spacing
- ▶ Screening/Scoping for final grid specification

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Geophysics Summary

▶ Preliminary ERT data

- fit of synthetic to simulated data provide high confidence in the resulting subsurface imagery
- heterogeneous material distribution
- time-dependent behavior

▶ Full Grid Deployment

- 120 SP electrodes
- 60 ERT/Complex Resistivity electrodes
- 300 Area flow behavior

▶ Next Steps

- Use borehole logs and depth to water to interpret layers in terms of lithology
- SP survey analyzed with hourly water level to identify groundwater flow field
- Complex resistivity for material property distribution
- Identification of permanent electrode locations

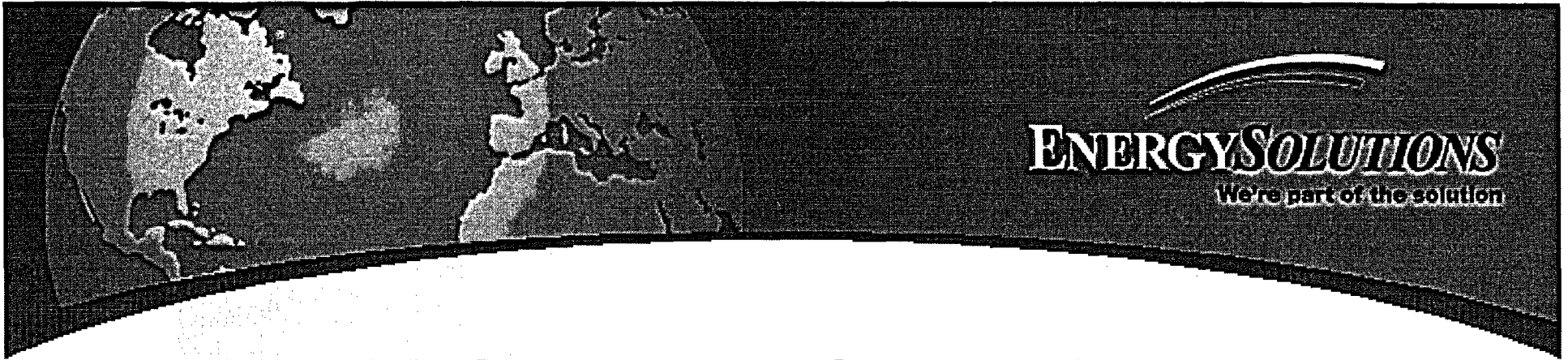


Role of Models in Demonstration, Compliance with Licensing Requirements

*presented to the ACNW
September 19, 2006
by Vernon Ichimura*

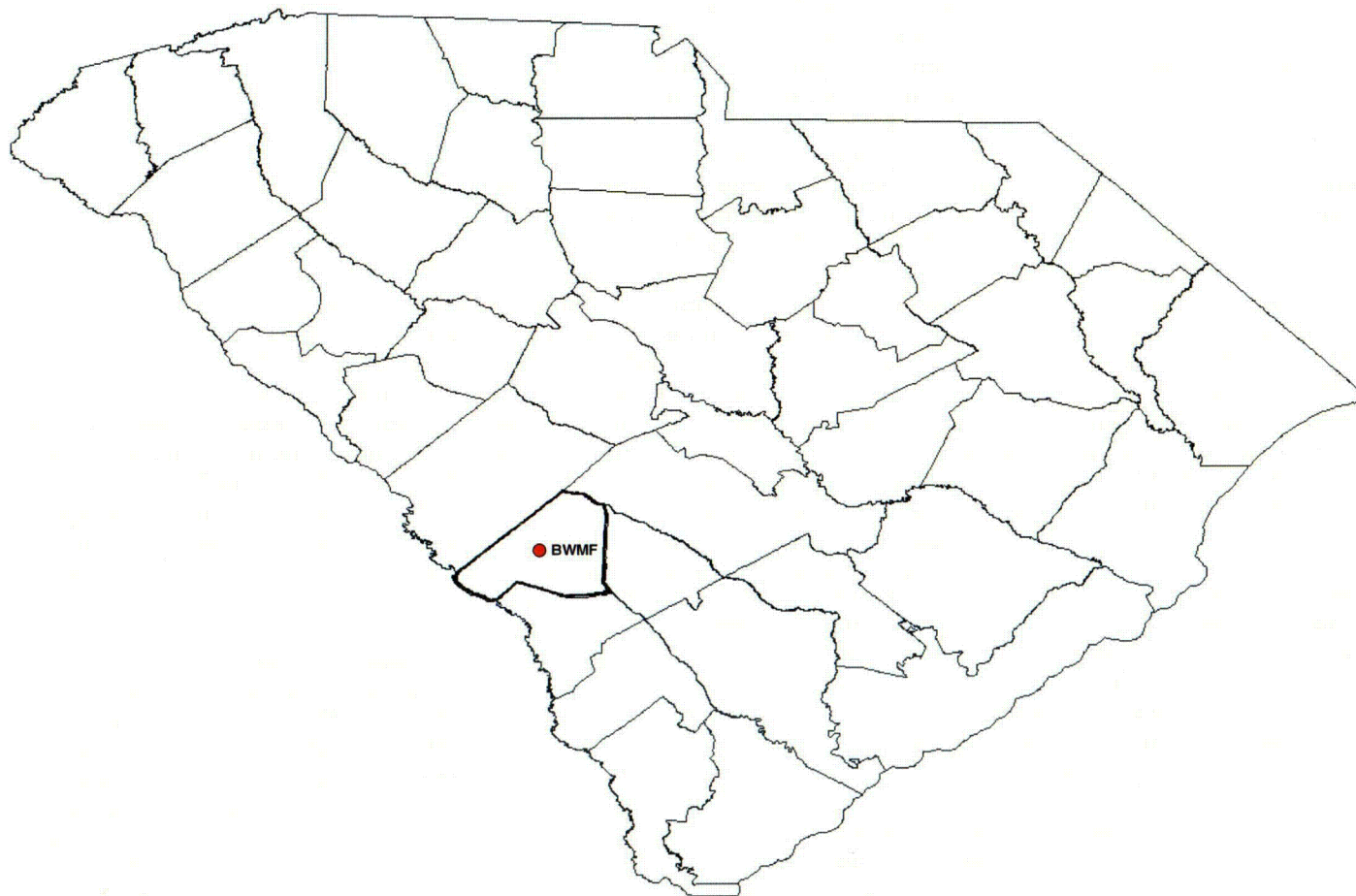
Overview

- Barnwell Disposal Site
- Review of Regulation
- Focus on Measurement
- Use of Models



Barnwell Disposal Site -- Summary

- Licensed to dispose LLRW in 1971
- Current license area is 235 Acres
- Approximately 12 million curies received
- After decay, approximately 3 million curies remain
- Current area used for disposal is 105 acres
- Approximate area remaining is 10 acres
- Approximate disposal volume is 28 million cubic feet
- Approximate disposal volume remaining is approximately 2 million cubic feet

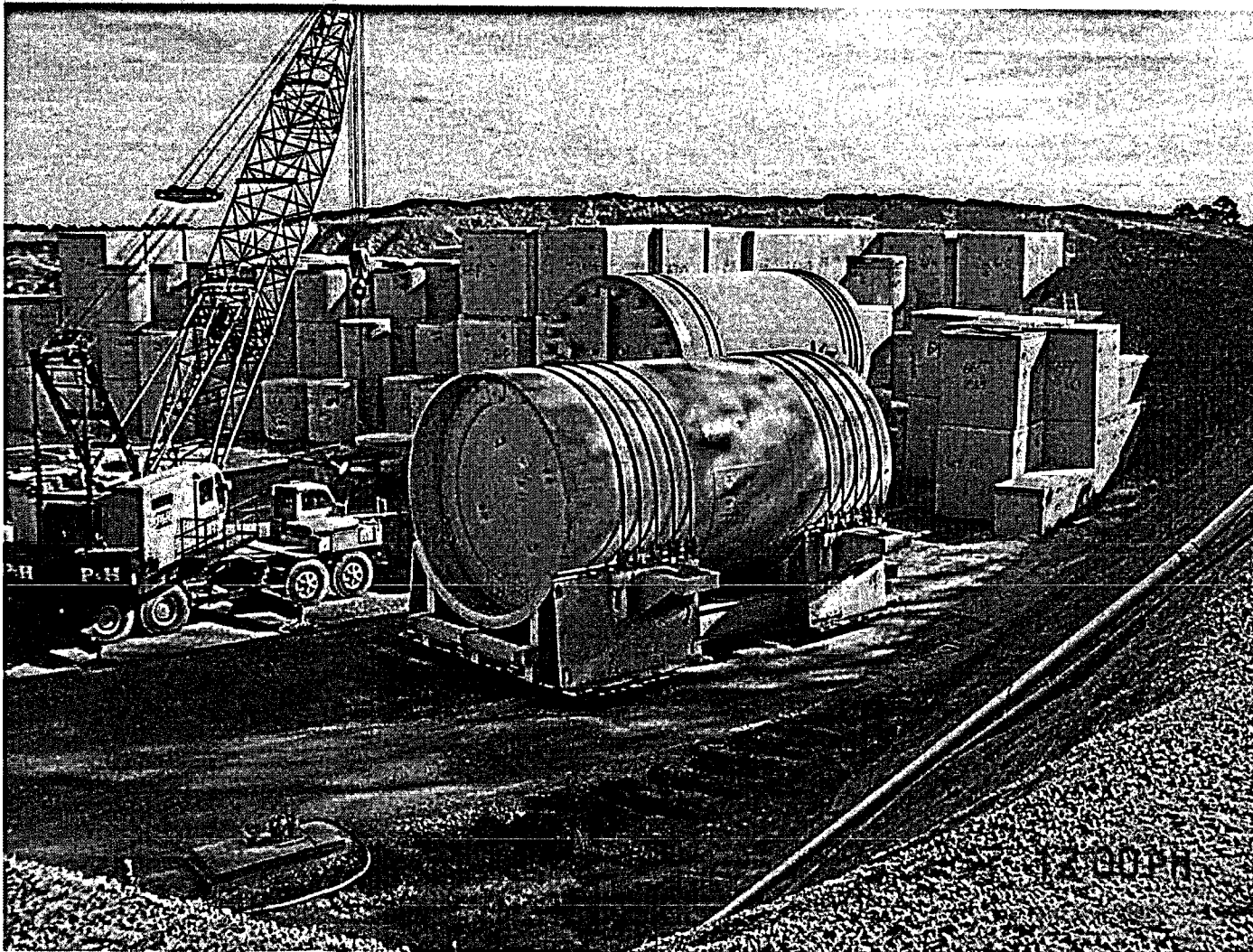






ENERGY SOLUTIONS

We're part of the solution



9/19/2006



Regulations

Demonstrate by measurement and/or model during operations and after site closure that concentrations of radioactive materials which may be released to the general environment in groundwater, surface water, air, soil, plants or animals will not result in an annual dose exceeding an equivalent of 25 millirems to the whole body, 75 millirems to the thyroid, and 25 millirems to any other organ of any member of the public.



Operational

- Real dose to workers
- In 2005 – Average Annual Dose to a Radiation Workers was 241 millirems

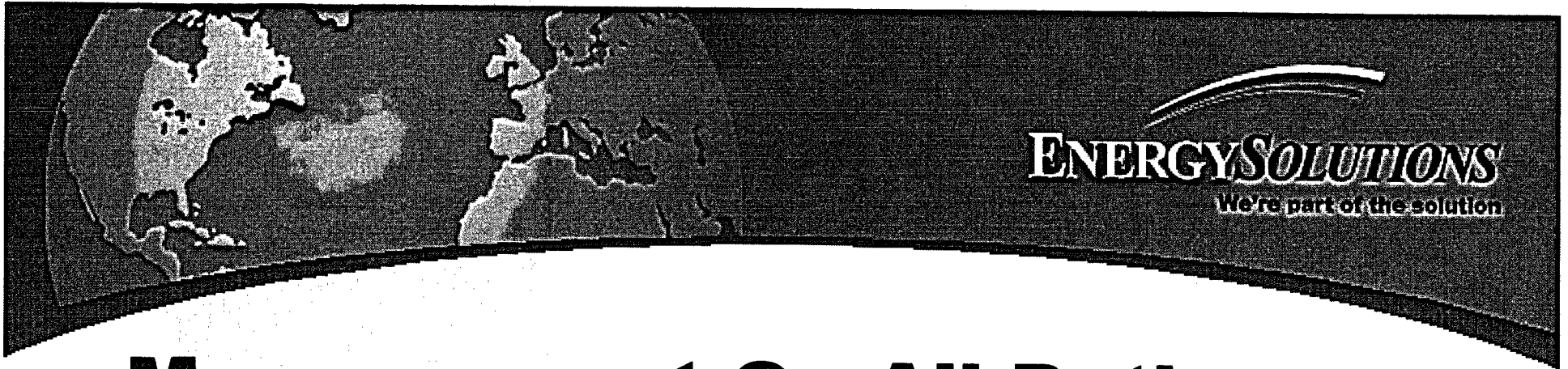
Environment

- Hypothetical dose to any member of the public
- In 2005 – Average Annual Dose to Public - - Negligible
- In 2005 – Average Hypothetical Dose by Groundwater/Surface Water at the Compliance Location is less than 5 millirems.



Focus on Measurement

- At location adjacent to waste disposal operations
- Around and in closed disposal trenches
- On the disposal site
- At boundary and compliance locations
- At off-site locations around the disposal site
- Distant from the disposal site - - for background evaluations



Measurement On All Pathways

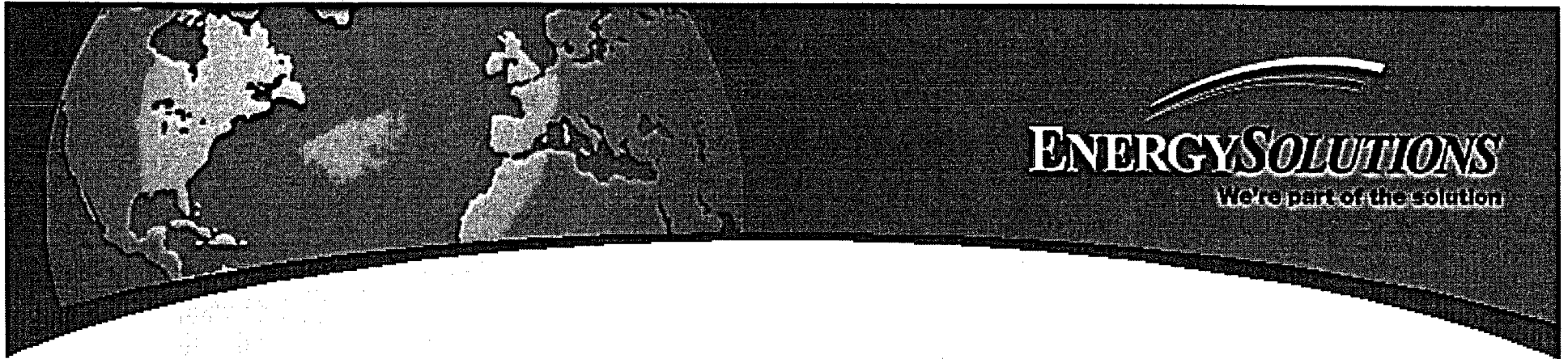
- Direct exposure
- Airborne
- Surface Water
- Soils
- Plants
- Groundwater

Use of Simple Models or Well Documented Models Which Have Been Checked

- Simple “calculator”, handbook, and analytical models - - - based on theoretical principles
- Commercial or public domain models
- Run validation
- Check model results with measurements
- Independent “peer-review” of model and projections

Use of Models - - Examples

- Estimate boundary dose rates due to disposal operations
 - What is the necessary shielding required for groups of waste packages and waste configurations
 - Simple inverse square law models and Microshield[®]
 - Verify with measurements



- Estimate radionuclide concentration at the site boundary in surface soil and surface water
 - Measurements of radionuclide concentration in soils
 - Erosion calculations and measurements
 - Runoff calculation and measurements
 - Estimate radionuclide concentration at the boundary
 - Verify with measurements

- Estimate radionuclide concentration at a compliance location in groundwater and surface water
 - Measurements of radionuclide concentrations
 - Measurements of hydraulic data
 - Perform groundwater flow and transport modeling
 - Verify with measurements

Roles of Models

- Models are needed to demonstrate compliance
- Models are simplification of reality and contain numerous assumptions
- Models must be checked with measurements
- Models should be updated as new information becomes available



Groundwater Monitoring in Support of License Termination at Yankee Nuclear Power Station

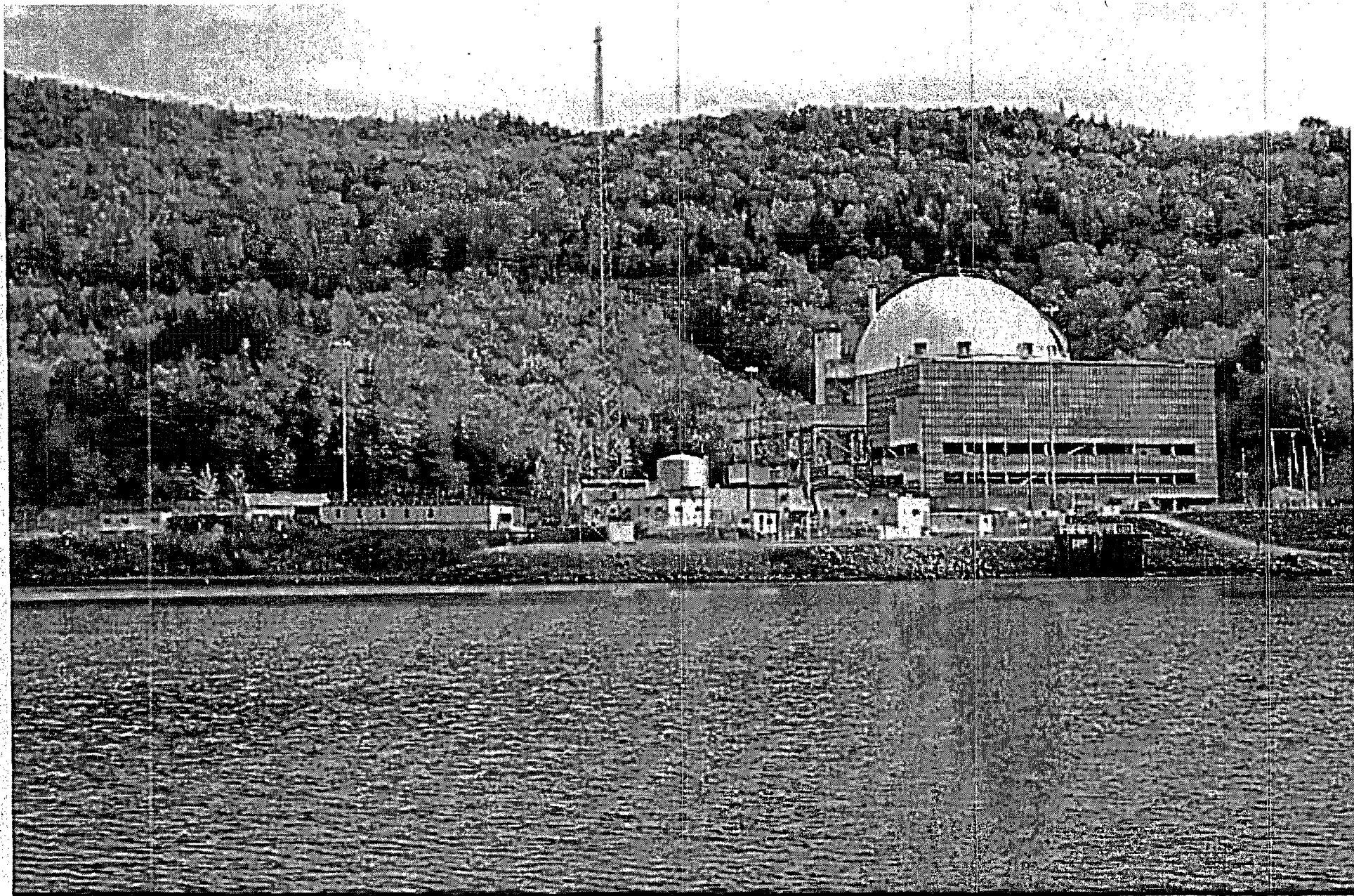
**Advisory Committee on Nuclear Waste
U.S. Nuclear Regulatory Commission
Rockville, MD, September 19 & 20, 2006**

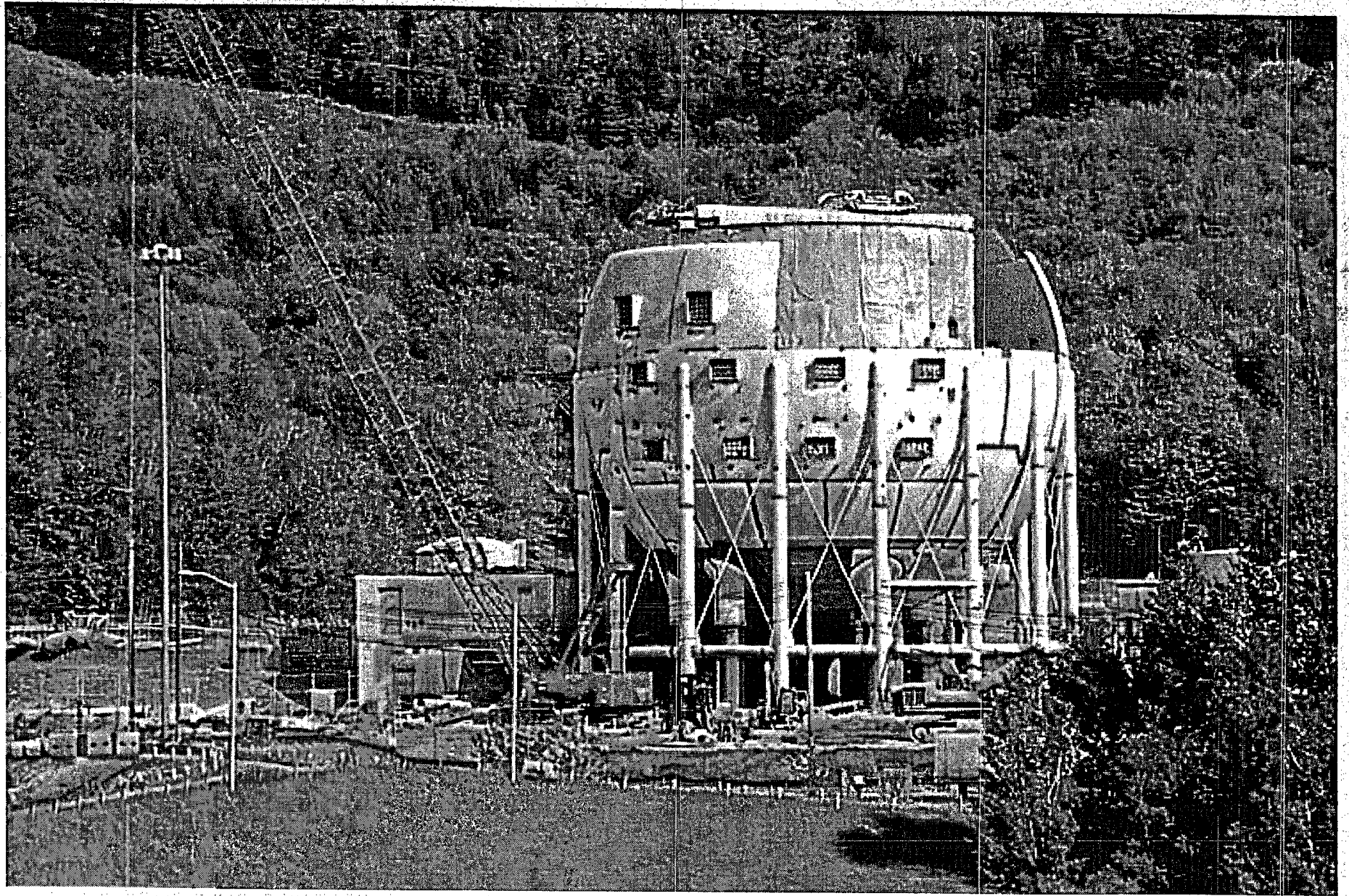
Dave Scott, Project Hydrogeologist, Radiation Safety & Control Services, Inc.
Greg Babineu, Yankee Atomic Electric Company
Eric Darois, CHP, Radiation Safety & Control Services, Inc.

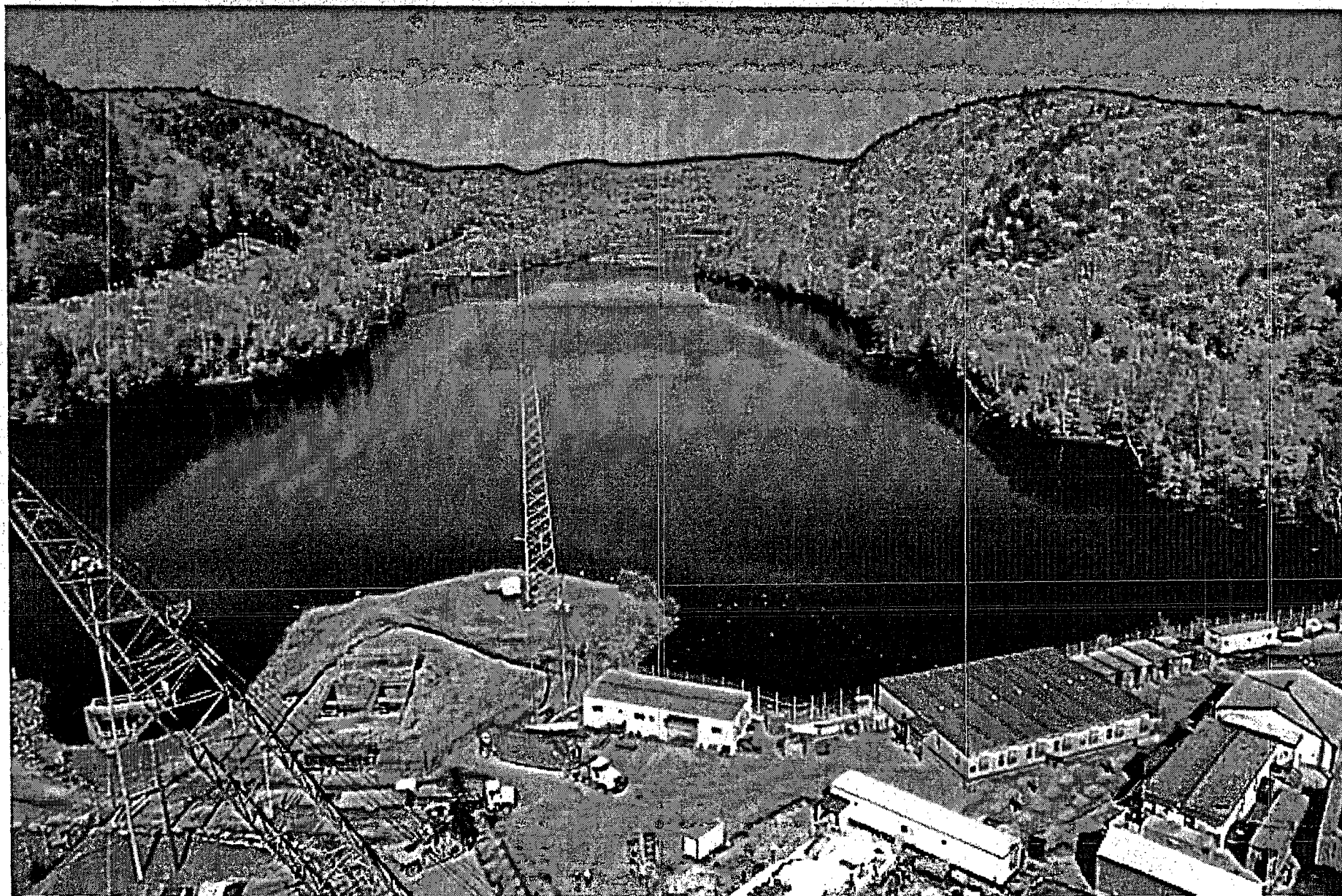


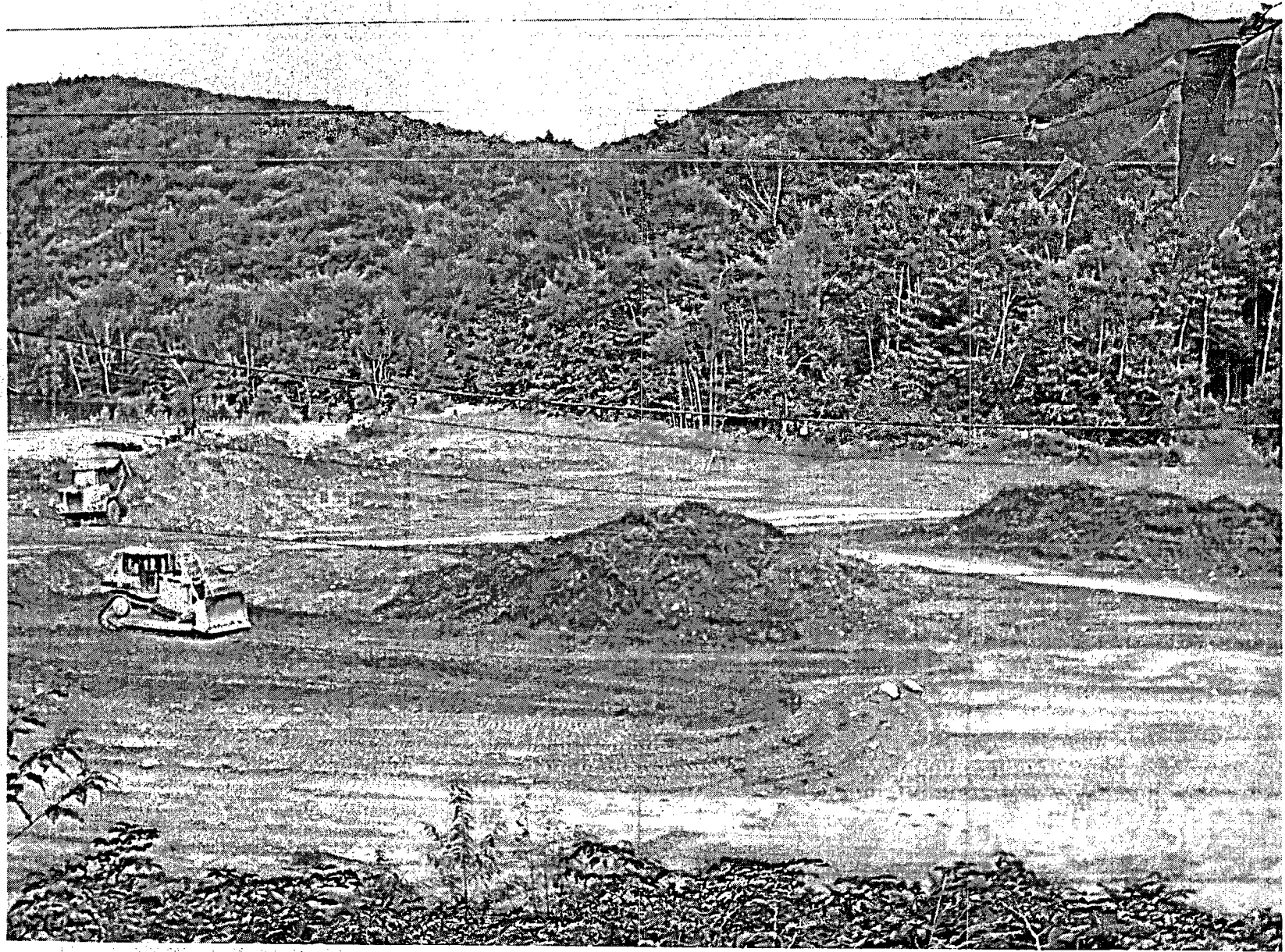
YR Operational History

- PWR, Operated from 1960 to 1992
- Built adjacent to Sherman Reservoir in the northern Berkshires using a Vapor Containment Design
- Initially 485 MWt, Upgraded to 600 MWt in 1963
- Permanently Ceased Operations in 1992
- Significant IX Pit Leak in 1963
- Fuel Clad for ~14 years was Stainless Steel
- During the period 1960-1980 the SFP did not have an interior stainless liner











Yankee Rowe Potential Groundwater Contaminating Events

- SFP Unlined From 1960 Until 1980
- IX Pit Leak First ID'ed in 1963; Repaired in 1965
- Outside Storage Of Contaminated Materials
 - Refueling Equipment
 - Waste
- Redistribution of Soil Contamination
 - RCA Snow Removal
 - Rain – Storm Drains
 - Wind
- RX Head Impact – Outside Soil Contamination
- Underground Drain Pipe Leak in Radwaste Warehouse



Criteria For License Termination

- All Pathways TEDE < 25 Millirem/yr (10 CFR 20.1402), and Residual Radioactivity ALARA
- H-3 Concentration in Resident Farmer's Well Less Than 20,000 pCi/L
 - Average yield of well serving family of four: 1323 m³/yr (0.665 gpm)
- Other GW Contaminants Less Than Limits Defined in LTP License Condition



Initial GW Monitoring Activities

- First 10 Monitoring Wells Drilled in 1993
- 24 Wells Added During '94, '97, '98 and '99
 - Virtually all in shallow outwash aquifer <30 feet deep
 - 18 in radiologically controlled area (RCA)
 - 5 in industrial area outside RCA
 - 3 outside industrial area
 - 8 in construction fill area, upgradient of RCA
- 2 Additional Monitoring Points
 - Sherman Spring (monitored since 1965)
 - Plant potable water well (bedrock)

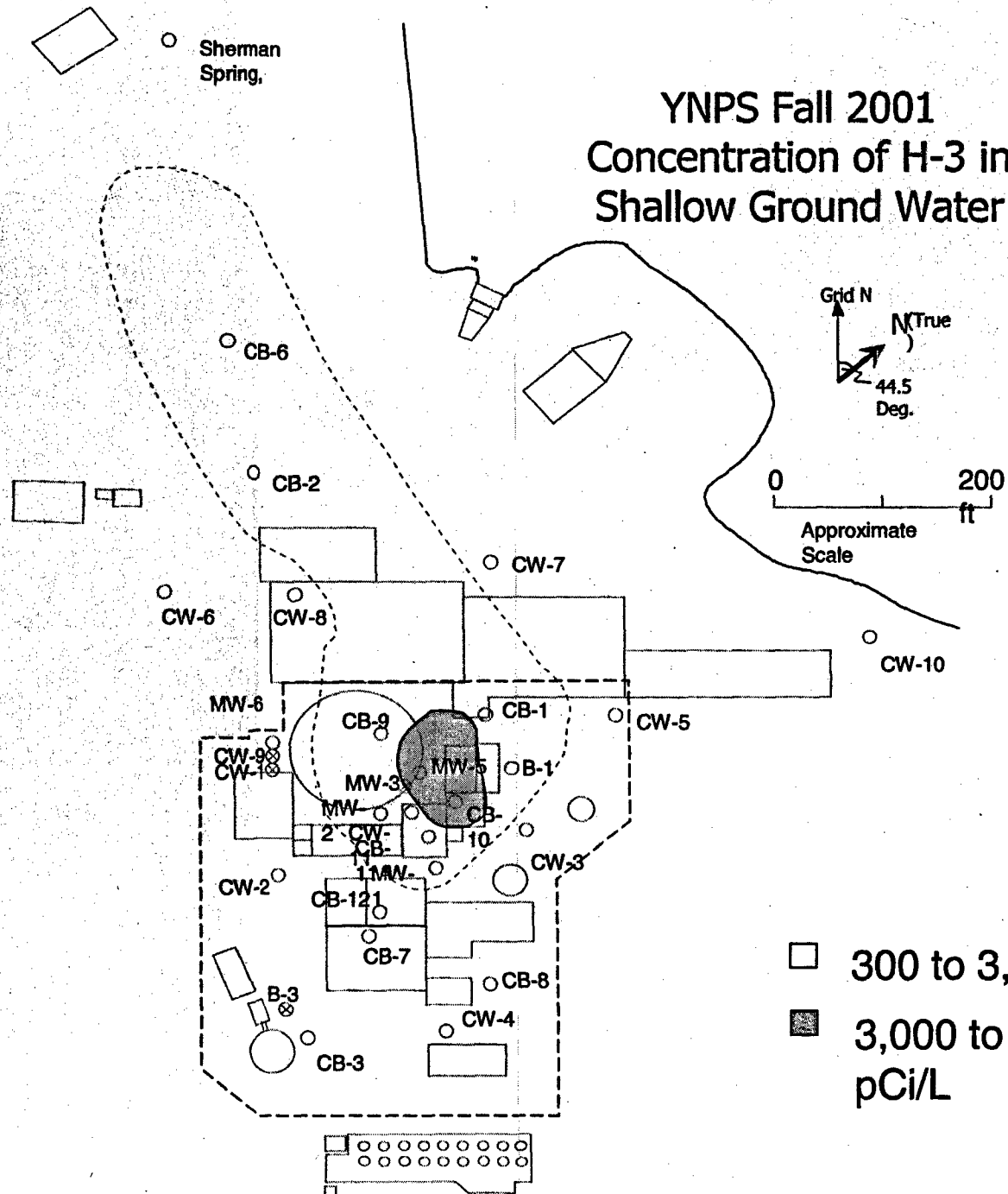


Initial GW Monitoring Activities (continued)

- Periodic Sampling and Analysis for:
 - Tritium
 - Gamma-emitters
 - Chemical constituents
- One Round of Analysis for Sr-90
- Identified Tritium Plume
 - Maximum concentration ~ 5,000 pCi/L
 - Extends downgradient from SFP/IXP



YNPS Fall 2001 Concentration of H-3 in Shallow Ground Water





Comprehensive GW Monitoring From 2003

- Evaluated Accumulated Historic GW Data
- Resulting Recommendations:
 - Drill additional wells
 - Fully characterize deeper aquifers beneath outwash, down to bedrock
 - Improve procedures for drilling, sampling & analysis
 - Define DQO/DQA
 - Begin use of roto sonic drilling, low-flow sampling, quarterly sampling
 - Standardize and expand list of radionuclide analytes to 22



Comprehensive GW Monitoring From 2003

- Established monitoring program that included:
 - Suites of radionuclide analytes determined by location, based on HSA and LTP
 - New locations for wells based on site geology
 - Intermediate depth sand lenses (30 -100 feet)
 - Bedrock (some as deep as 300 feet)
 - Multiple wells at same location for vertical profile
 - Frequency of monitoring that will adequately measure changes in GW quality



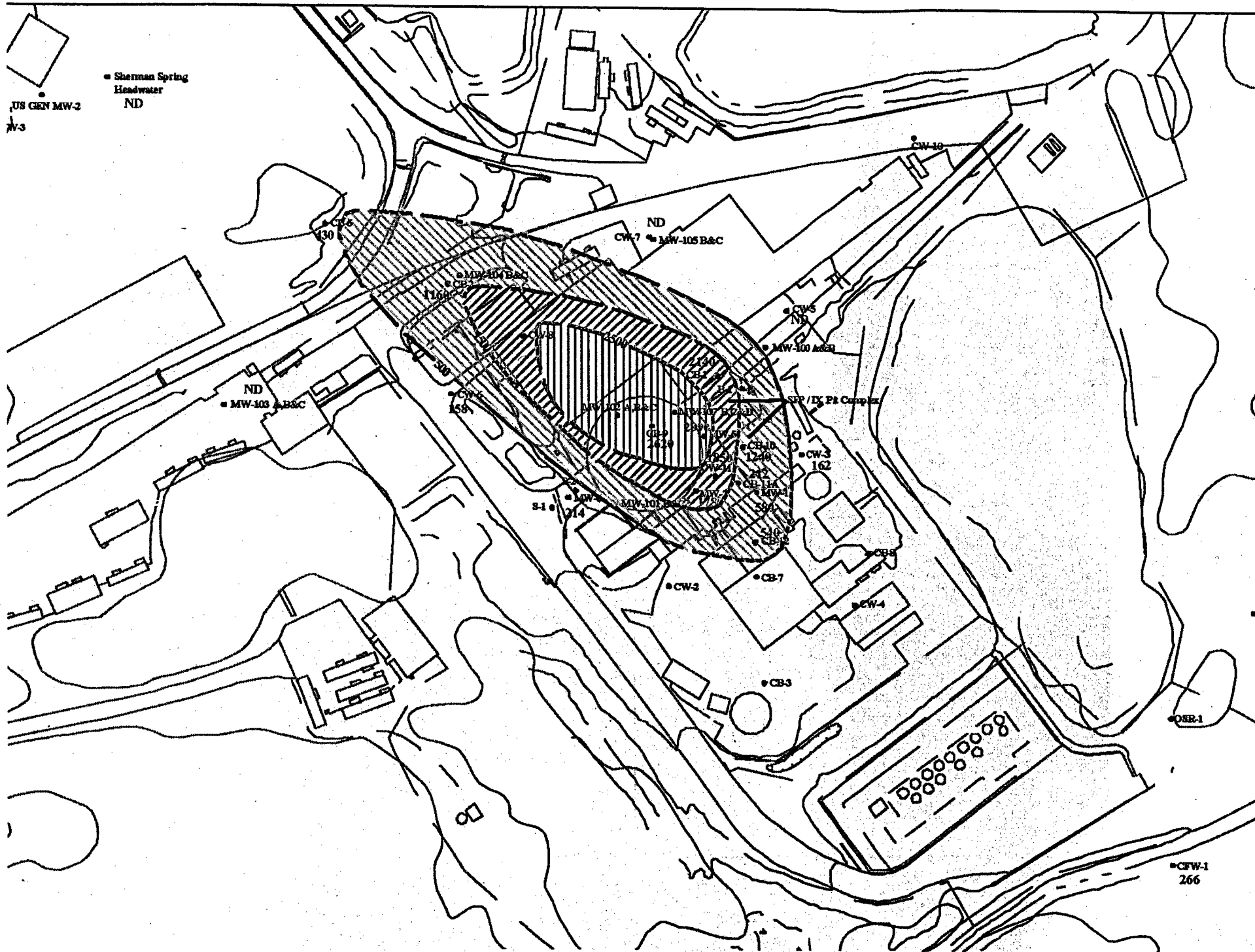
2003 MW Drilling Program

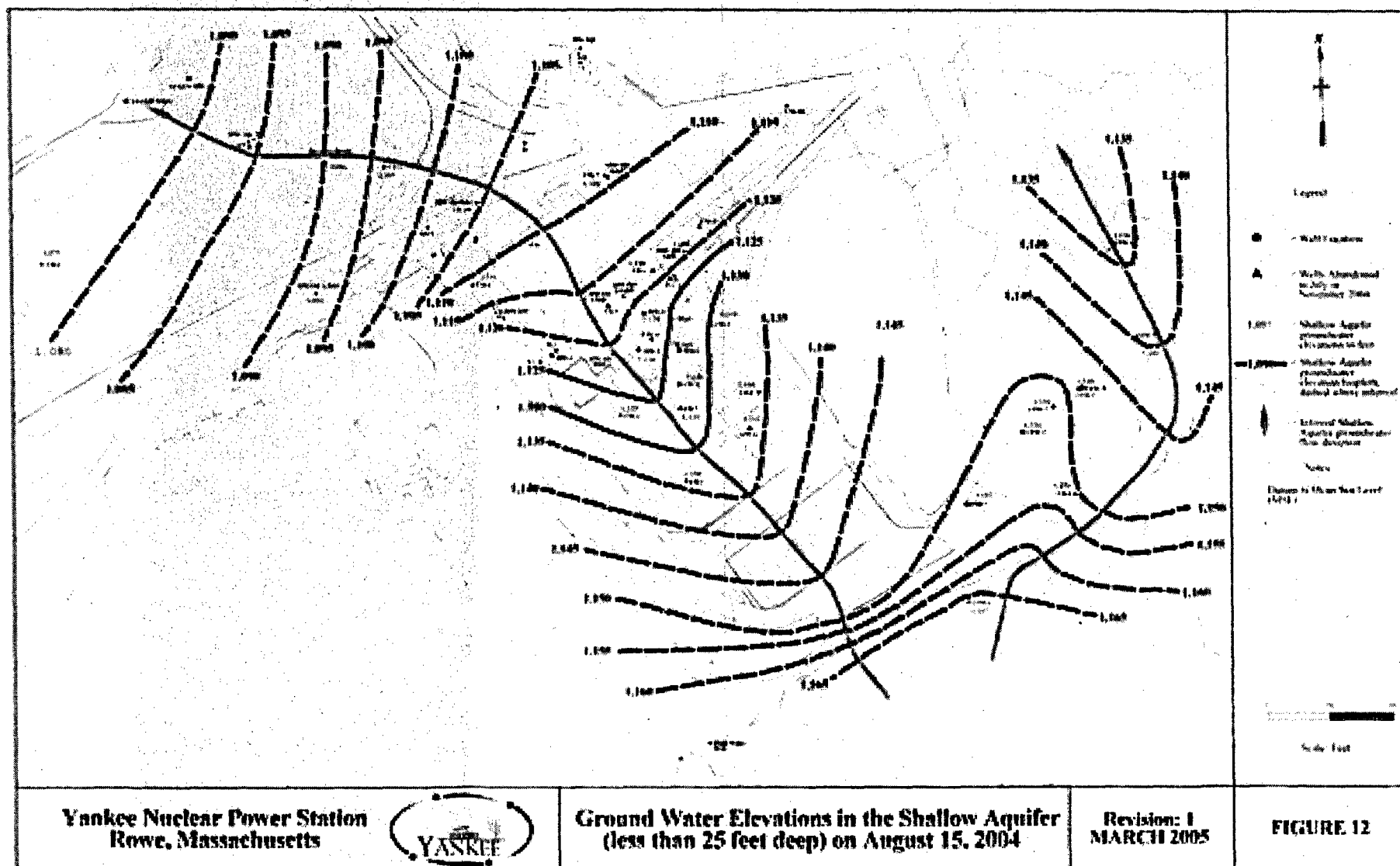
- 17 Wells Installed by Rotosonic Method
 - “Telescoped” up to 4 drill casings to properly isolate multiple aquifers
 - Characterized complex stratigraphy
 - Determined vertical distribution of tritium
- Explored Entire Thickness of Sediments and Shallow Bedrock
 - 2 wells into shallow outwash aquifer
 - 8 wells into deeper sand lenses interlayered within underlying lodgement till
 - 7 wells into bedrock
- Maximum depth of 295 feet



Results of 2003 Investigation

- Tritium only plant-related radionuclide in GW
- One H-3 plume in shallow (outwash) aquifer
 - Maximum concentration ~ 3,500 pCi/L
 - Aligned with direction of shallow GW flow (NW)
- A second H-3 plume in deeper sand lenses
 - Maximum concentration ~ 45,000 pCi/L
 - Direction of deeper GW flow toward Deerfield River
- H-3 in one bedrock well ~ 5,000 pCi/L







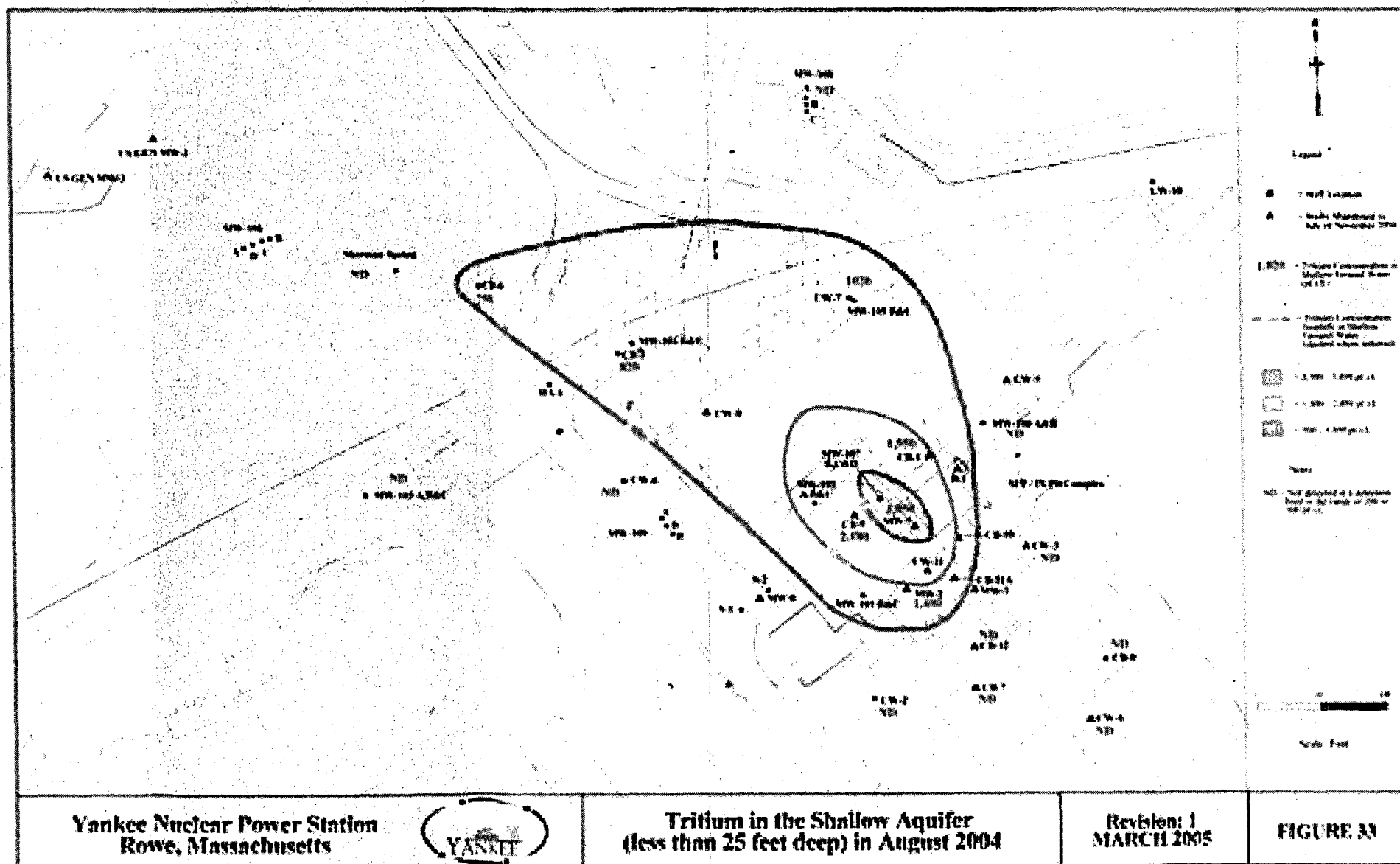
2004 MW Drilling Program

- 10 Additional Monitoring Wells Installed by Rotosonic Method
 - 2 into shallow outwash aquifer
 - 5 into deeper sand lenses interlayered within underlying lodgement till
 - 3 into bedrock
- Well Locations Chosen to Bound the Shallow and Deeper H-3 Plumes
- Studied Interconnectivity Between Aquifers by Monitoring GW Levels With Data-Logging Pressure Transducers



Groundwater Flow Characteristics

- Flow in Shallow Aquifer Relatively Fast (1 to 2 feet per day, or $K \sim 5$ ft/day)
- Net Flow Rate in Deeper GW is Much Slower – Controlled by Discontinuous Sand Lenses Within Lower Permeability Matrix of Lodgement Till





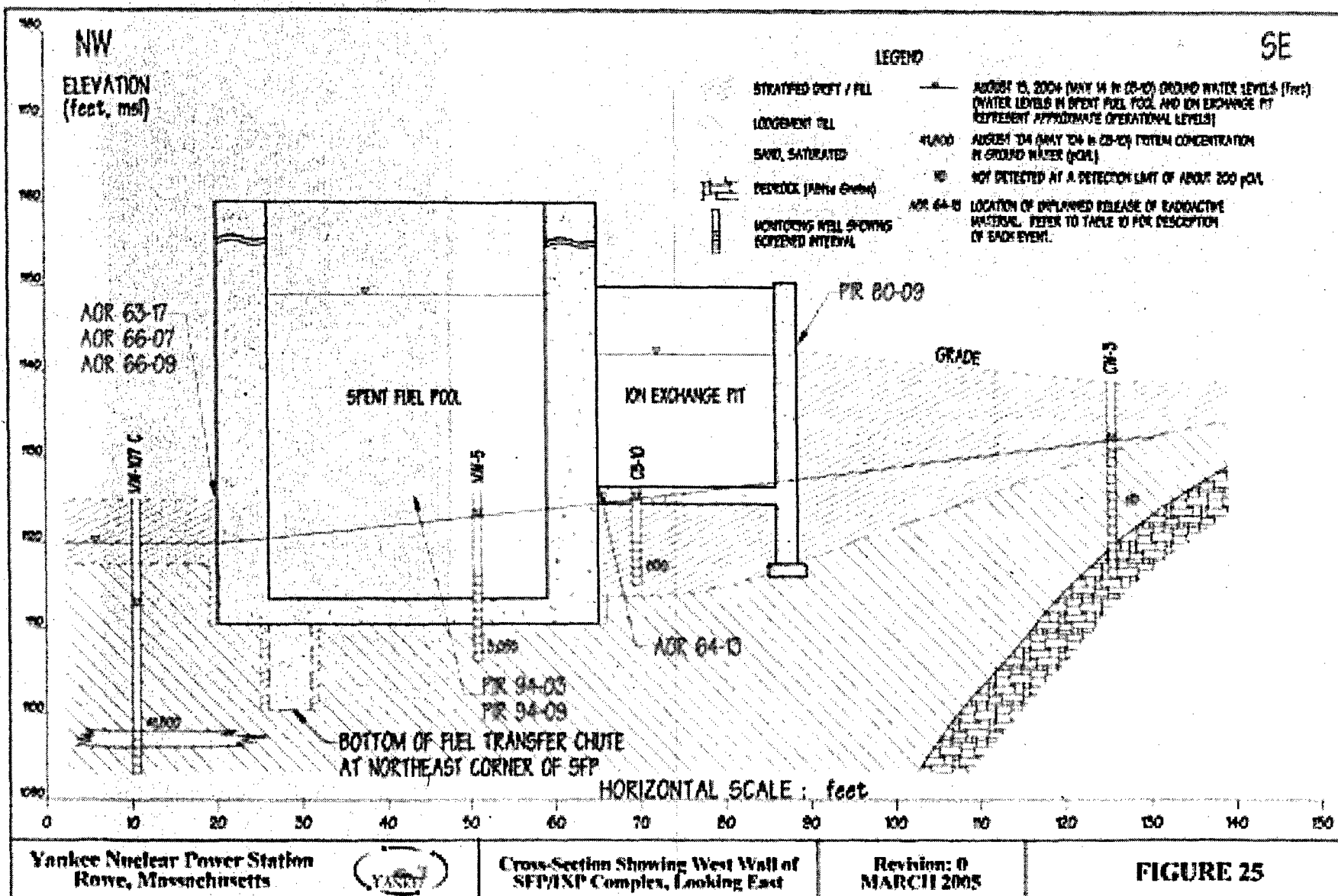
Source of Tritium Plumes

- Primary Source is the SFP/IX Pit Complex:
 - Maximum H-3 concentration occurs close to SFP/IX Pit in both shallow and deeper aquifers
 - IX Pit is known to have leaked ~ 1963
 - Repaired early in 1965
- REMP Monitoring Detected Tritium in Sherman Spring, 550 feet Downgradient of SFP/IX Pit
 - Peaked ~ 7.2×10^6 pCi/L in Dec 1965, after IXP repaired
 - Declined continuously (<200 pCi/L since '93 except for spike during demolition in 2005)
- IX Pit Emptied in 1995, Demolished in 2005
- SFP Emptied in 2003, Demolished in 2005



Contaminant Transport Mechanisms

- Tritium Entered Deeper GW Along Deep Foundations and Piping
 - Downward flow potential in vicinity of SFP/IXP, shown by multiple-depth well clusters
- H-3 Became “Trapped” in Deeper Sands and Slowly Diffuses into Shallow Aquifer
 - This condition may sustain the low-concentration shallow plume, which otherwise may have attenuated

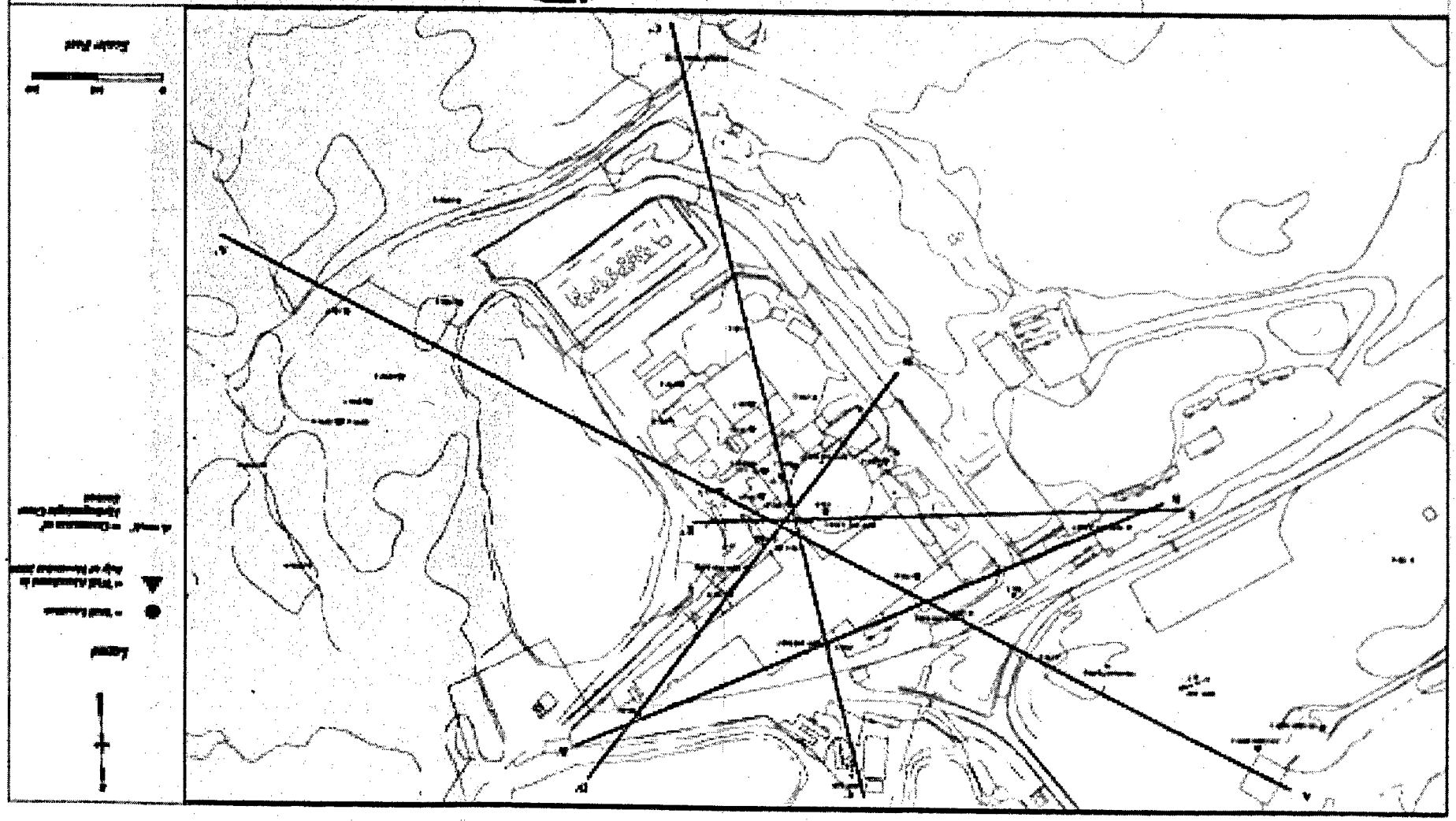


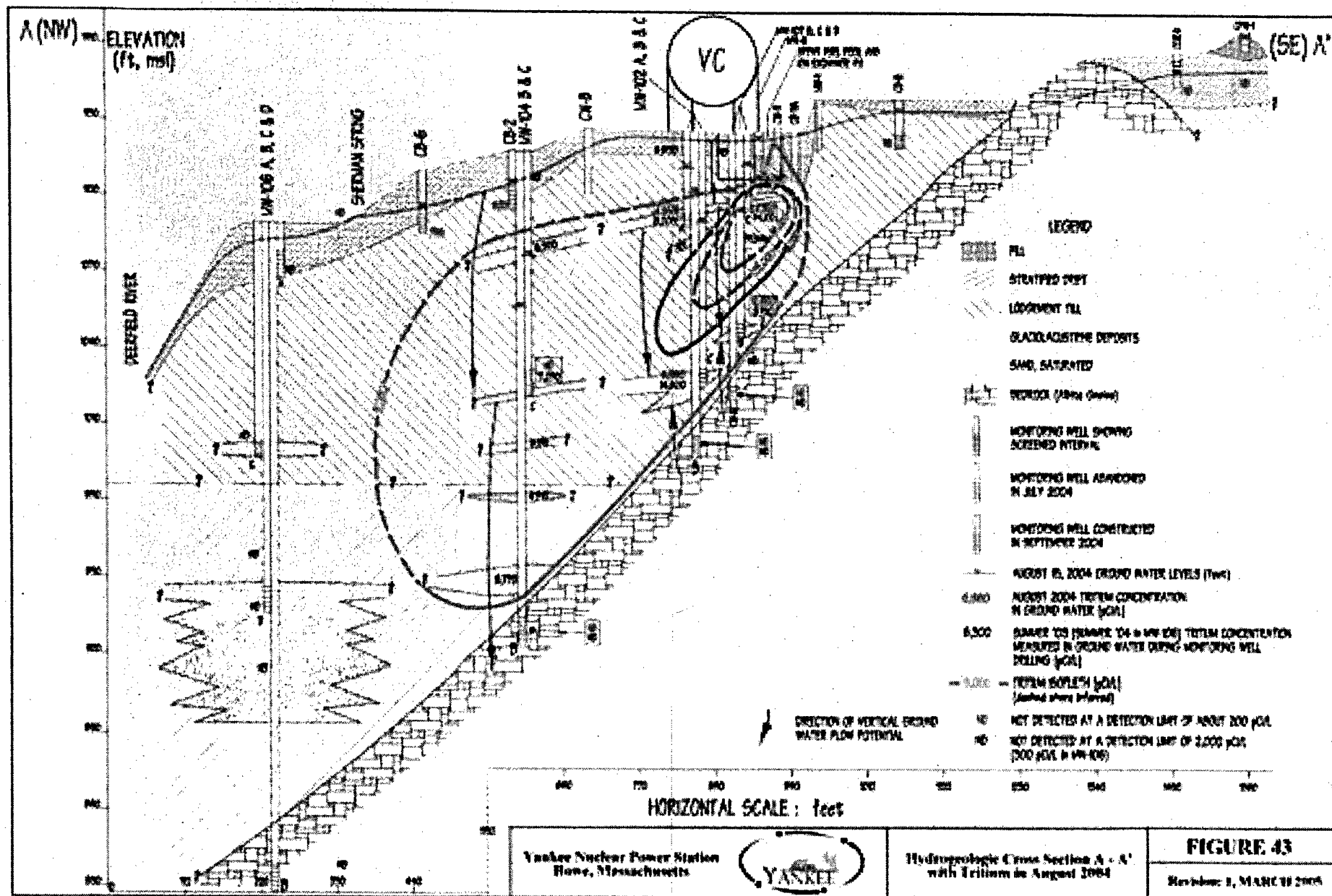
Yankee Nuclear Power Station
Location of Cross Sections



Date: February 2005
Revision: 1

Figure 7

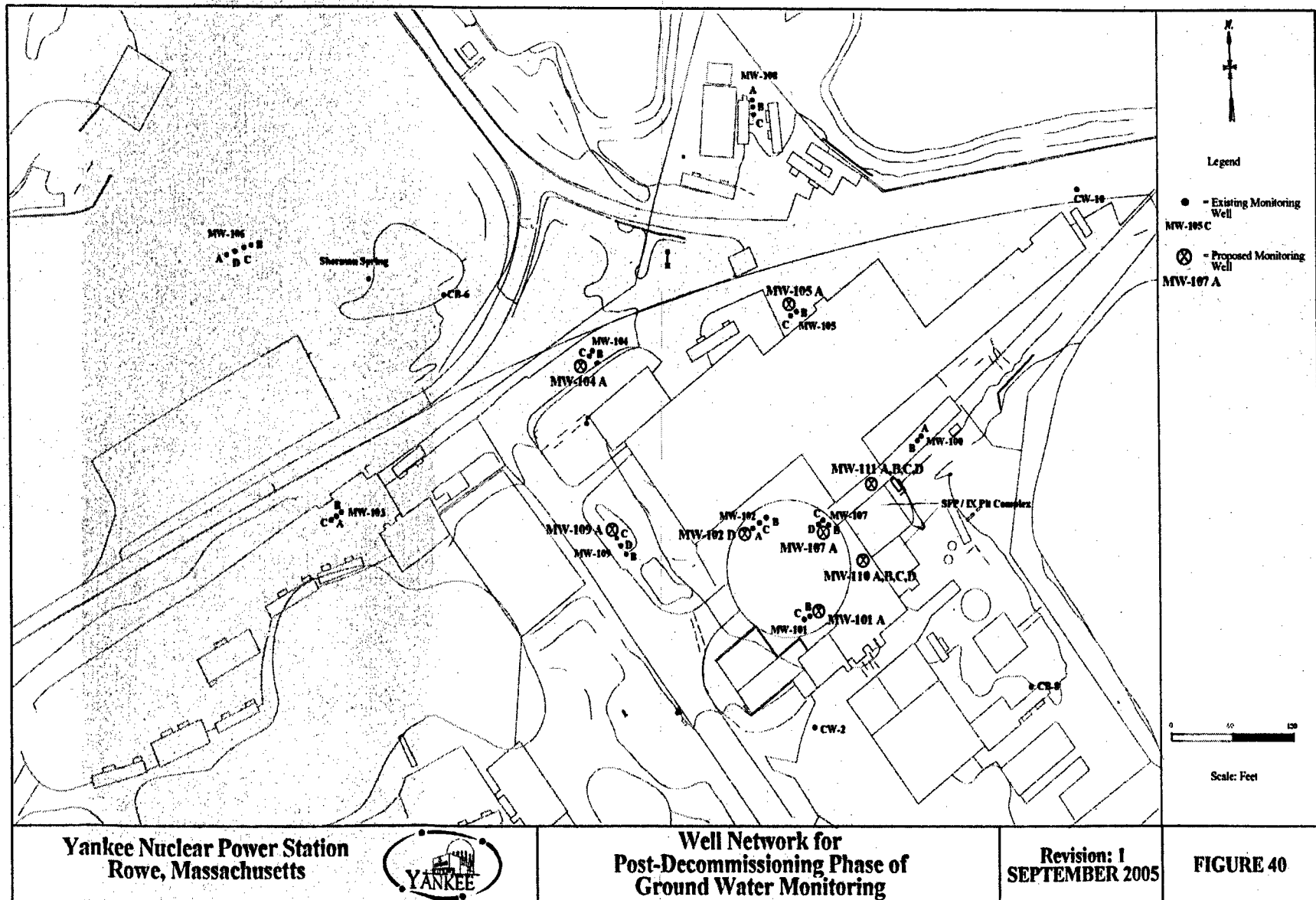






2006 MW Drilling Program

- 17 Additional Wells Drilled by Rotosonic Method
- 3 Multi-Depth Well Clusters Drilled in Key Locations:
 - At IX Pit Leak
 - Adjacent to Lowest Part of SFP Foundation
 - Downgradient of Septic Leach Field
 - To Confirm Plume Source and Absence of Additional Radionuclides in GW Other Than H-3
 - To Better Define Interconnectivity of Aquifers
- 2 to Bound the Sand Lens with Highest H-3
- 6 Shallow Wells to Replace a Few Abandoned to Facilitate Plant Demolition





Preliminary Results of Ongoing 2006 Investigation

- H-3 Still the Only Plant-Related Nuclide
- Drilling Results Confirm Sand Lenses in Deep Till are of Limited Extent
- Pumping Tests Conducted to Determine:
 - Hydrogeologic Parameters (K, S) for Key Lenses
 - 24-Hr Constant Rate Test in Well With Highest Tritium
 - Hydraulic Connection Between Sand Lenses
 - 2-Hour Pressure Transient Tests in 12 Selected Wells
 - Pressure Transducers Monitor WL in Nearby Wells



Preliminary Results of Ongoing 2006 Investigation (con't)

- Numerical Fate and Transport Computer Model Under Development
 - Will Incorporate:
 - Stratigraphic Model From Drilling Results
 - Water Level Measurements With PXDs
 - Groundwater Sample Analysis Results
 - Pumping Test Results
 - To Validate Site Conceptual Model
 - To Predict H-3 Concentrations at Compliance Point
 - To Demonstrate Compliance with Criteria for License Termination



Yankee Rowe Lessons Learned

- The Rowe Site has Multiple Aquifers
- Contamination can Migrate Through Multiple Aquifers to Depths >100 feet.
- Hydrogeologic Investigation is an Iterative Process
- Important to Develop a Hydrogeologic Conceptual Site Model:
 - To Aid Well Placement
 - To Understand Contaminant Transport
 - To Define Aquifer Characteristics
- Long Term Data Trends Are Important
 - Allow Bias Detection
 - Identify Seasonal Fluctuations
 - Identify New Contaminant Releases



Yankee Rowe

Lessons-Learned (continued)

- **Water-Level Monitoring is Instructive**
 - May Demonstrate Connection or Isolation of Aquifers
 - Useful for Calibration of Numerical Model
- **Early Investigations at YNPS Not Sufficiently Rigorous**
 - MWs not deep enough
 - Little Regulatory Involvement
- **Involve All Stakeholders**
- **Analyze for Wide Suite of Radionuclides**
- **Include Non-Rad Constituents for Site Closure**



Response to Selected ACNW Working Group Focus Questions

- **Q1. Why are GW compliance monitoring data not used to enhance confidence in numerical models after site characterization and licensing is complete**
 - **Regarding operating power stations: GW characterization during plant design and construction was not sufficiently detailed to support contaminant fate & transport models**
 - **GW monitoring methods were in their infancy when the last power station was built (early 1970s)**
 - **Rigorous GW investigation should occur during plant construction with wells drilled near and downgradient from key sources of primary water:**
 - **Spent Fuel Pool**
 - **Refueling Water Storage Tanks**
 - **Condensate Tanks**

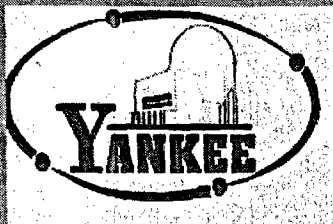


- Data from the initial detailed investigation can be used to:
 - Build a numerical model
 - Respond to contaminant releases more expeditiously (stratigraphy, GW flow directions and contaminant flow paths already known)
- Long-term GW monitoring data used to:
 - Detect contaminant releases
 - Refine numerical model - change in state variables measured over time used to improve model calibration
 - Hydraulic head (water levels)
 - Water temperature
 - Tidal influence
 - Surface water stage
 - Contaminant concentration temporal trends
 - » Tritium or other radionuclides
 - » Hydrocarbons, solvents and degradation products
 - » Inorganic constituents: chlorides, boron



Response to Selected ACNW Working Group Focus Questions

- **Q6. New Methods and Analytical Tools That Should be Pursued:**
 - GW age determination by measuring the ratio of ^3H to ^3He may improve calibration of models of some GW systems
 - Aid definition of GW flow paths
 - Identify contaminant transport zones
 - Soil-gas surveys of ^3He concentrations can be useful for delineating shallow tritium plumes



Questions?